Dissolved organic carbon and nitrogen leaching from Scots pine, Norway spruce and silver birch stands in southern Sweden

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1 Abstract

2 The effects of three common tree species – Scots pine, Norway spruce and silver birch – on leaching of 3 dissolved organic carbon and dissolved nitrogen were studied in an experimental forest with 4 podzolised soils in southern Sweden. We analyzed soil water collected with lysimeters and modeled 5 water fluxes to estimate dissolved C and N fluxes. Specific UV absorbance (SUVA) was analyzed to get 6 information about the quality of dissolved organic matter leached from the different stands. Under the 7 O horizon, DOC concentrations and fluxes in the birch stands were lower than in the spruce and pine 8 stands; annual fluxes were 21 g m⁻² y⁻¹ for birch and 38 g m⁻² y⁻¹ and 37 g C m⁻² y⁻¹ for spruce and pine, 9 respectively. Under the B horizon, annual fluxes for all tree species ranged between 3 and 5 g C m⁻² y⁻¹, 10 implying greater loss of DOC in the mineral soil in the coniferous stands than in the birch stands. We 11 did not find any effect of tree species on the quality of the dissolved organic matter, as measured by 12 SUVA, indicating that the chemical composition of the organic matter was similar in leachates from all 13 three tree species. Substantial amounts of nitrogen was leached out of the soil profile at the bottom of 14 the B horizon from the pine and birch stands, whereas the spruce stands seemed to retain most of the 15 nitrogen in the soil. These differences in N leaching have implications for soil N budgets. 16 17

18 Keywords

19 Dissolved organic matter; tree species; dissolved organic carbon; forest soil; nitrogen; DOC quality

20

21 **1. Introduction**

22 The dominant tree species in a forest may have fundamental effects on soil properties, both biological, 23 chemical and physical (e.g. Binkley and Giardina, 1998). Three different tree species constitute the 24 majority (>90%) of Swedish forests: Norway spruce (Picea abies) with 41%, Scots Pine (Pinus 25 sylvestris) with 39% and Birch (Betula pendula/B. pubescens) with 12% of standing volume 26 (Anonymous 2010). A large majority of Scandinavian forests is managed and re-planting after harvest 27 is common practice. The choice of tree species is an important forest management decision, which may 28 also have implications for soil and ecosystem C and N budgets. In the project reported here, we have 29 studied the effects of the three dominant Swedish tree species on dissolved C and N.

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Dissolved organic carbon (DOC) may at first seem to be a negligible part of the C budget in forest soils.
 The DOC flux usually corresponds to a C flux of about two orders of magnitude or less compared to

total gross ecosystem fluxes (Gielen et al. 2010; Kleja et al. 2008). It is however becoming increasingly
obvious that such comparisons are not relevant and DOC should instead be compared to net ecosystem
fluxes (Gielen et al. 2010; Siemens 2003). For example, Luyssaert et al. (2010) estimated that the

36 average net European forest C sink is 75 ± 20 g C m⁻² y⁻¹, of which 22 g C m⁻² y⁻¹ was accumulating in the

soil. By comparison, the flux of C leaving the root zone as DOC from temperate forests is typically
about 1-20 g m⁻² (summarized by Michalzik et al. 2001) and DOC leaching may thus play a significant

about 1-20 g m⁻² (summarized by Michalzik et al. 2001) and DOC leaching may thus play a significant
 role in the C budget of forest soils (Gielen et al. 2010). Furthermore, DOC also plays an important role

40 in the internal C cycling in soils, as DOC is transported from the top soil to deeper parts of the soil,

41 where it is retained by physico-chemical processes (Kalbitz et al. 2000). This leads to stabilization and

42 thereby to a significant increased residence time of C in the soil (Kalbitz and Kaiser, 2008).

43

44 In this study, we present data on how birch, spruce and pine affect the quantity and quality of 45 dissolved organic matter (DOM) leaching through the soil. Numerous studies of DOC transport in soils 46 under different tree species have been made, but there are only few systematic investigations of the 47 effects of the three dominant Scandinavian tree species in the literature. In previous studies of effects 48 of birch, spruce and pine in Finland (Kiikkilä et al. 2006; Smolander and Kitunen, 2002; Suominen et 49 al. 2003), higher concentrations of water extractable organic matter (WEOM) per gram of carbon were 50 found under birch and spruce than under pine in the O horizon, but no differences were found 51 between tree species for mineral soil samples. It can however be questioned how representative those 52 studies are for field conditions, as water extracts were used instead of soil water collected with 53 lysimeters. From other studies it has been concluded that WEOM data should be interpreted with 54 caution as concentrations and quality may differ from actual dissolved carbon transported through the 55 soil and collected with lysimeters (e.g. Fröberg et al. 2003). Here we present data based on collection 56 of soil water with lysimeters. We are only aware of one previous study where lysimeters have been used 57 to study DOC leaching from spruce, pine and birch. In that study, performed in northern Finland, 58 there were higher DOC concentrations from pine stands than from birch stands, with DOC 59 concentrations from spruce being intermediate (Lindroos et al 2011). Observations of differences in 60 soil properties between birch and pine or spruce stands (e.g. Hansson et al. 2011) made us hypothesize

61 that concentrations and fluxes of dissolved organic carbon would differ between stands dominated by 62 spruce, pine and birch. In particular, we expected to find lower DOC fluxes under birch than under 63 pine and spruce. As litter chemistry differs between the tree species, we also expected to find 64 significant differences in DOC quality between birch and the coniferous tree species.

65

66 The effects of nitrogen leaching by different tree species have been more intensively studied than the 67 effects on DOC. It is well established that the input of N via deposition is greater in high-leaf-area 68 conifers than low-leaf-area deciduous forests (Gundersen et al. 2009). It is not clear however, if this 69 also results in larger N leaching from coniferous stands. The effects of tree species on N leached from 70 stands dominated by difference tree species are often confounded, as different species tend to grow on 71 soils with differing fertility (Kristensen et al. 2004). In this study, performed in an experimental forest, 72 there are no such effects of initial differences in soil fertility. Hansson et al. (2011) reported, for the 73 same plots used here, that soil N stocks were larger in the spruce and pine stands than in the birch 74 stands. They suggested that difference in N leaching was one contributing reason for the differences in 75 soil N stocks. We therefore also wanted to test the hypothesis that there are differences in dissolved N 76 transport between the tree species. 77

78 **2. Methods**

79 **2.1. Site description**

80 The study area is located in the Tönnersjöheden Experimental Forest in south-west Sweden (56°40-81 41'N, 13°03-06'E). Mean annual air temperature is 6.4°C and mean annual precipitation 1053 mm. 82 Mean annual N deposition during 2005-2008 was about 11 kg N ha⁻¹ y⁻¹ (Karlsson et al. 2010). The 83 parent material is of glaciofluvial origin and the soils show signs of podzolisation, but are weakly 84 developed and may be classified as either podzols, arenosols or regosols. The experimental design 85 included stands of three different dominant tree species: Norway spruce (Picea abies (L) Karst.), Scots 86 pine (*Pinus sylvestris* L.) and silver birch (*Betula pendula* Roth), replicated in a block design (n=3, n=3)87 except for birch where n=2). The present stands were established in 1951-1963. Plot size ranged from 88 720 to 1296 m². More details about the study sites can be found in Hansson et al. (2011).

89

90 2.2. Soil solution sampling and analysis

91 In each plot, 3 zero-tension lysimeters, (30 cm squares, made of plexiglass and a polyethylene net) 92 were installed under the O horizon. Under the A horizon, at 10-16 cm depth from the soil surface, 2 93 Prenart Soil Disc lysimeters (Prenart Equipment Aps, Frederiksberg, Denmark) were installed per 94 plot. Under the B horizon, at 42-50 cm depth from the soil surface, 2 Prenart Super Quartz lysimeters 95 (Prenart Equipment Aps, Frederiksberg, Denmark) were installed. Water was collected monthly and 96 pooled per plot and horizon prior to DOC and total dissolved nitrogen (DN) analysis. O horizon 97 samples were filtered before analysis (0.2 or 0.45 µm). For A and B horizon samples, no further 98 filtration after collection with Prenart lysimeters was made, as they did not contain particulate organic matter. Analyses of DOC and DN were made with a Shimadzu TOC-VCPH and Shimadzu TNM-1,
 respectively. In addition specific UV absorbance was measured at 260 nm (Perkin Elmer Lambda 11)

- $101 \qquad \text{and pH was measured on all pooled samples. Specific ultraviolet absorbance at 260 nm (SUVA_{260}) was$
- 102 calculated as absorbance divided by DOC concentration.
- 103

104 **2.3. Water flux modeling**

105 The water fluxes were simulated using the CoupModel (Jansson and Karlberg, 2004). The driving 106 climate variables were daily sum of precipitation, daily averages of air temperature, air humidity, wind 107 speed and solar radiation. These variables, with the exception for wind speed, were provided from the 108 nearby Simlångsdalen. Wind speed was taken from the SMHI meteorological station at Ullared. 109 Further, precipitation was adjusted by +7% for the aerodynamic error in precipitation measurements.

110

Soil hydraulic parameters were estimated for each plot by using the measured soil texture and pedofunctions as proposed by Rawl and Brankensiek (1989). The leaf area indices (including understory) were measured in each plot during summer 2010. The root densities were estimated from the measured root biomass. The other stand parameters were taken from Alavi (2002), Gärdenäs & Jansson (1995) and Richardson & Berlyn (2002) for spruce, pine and birch respectively.

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117 2.4. Statistical analysis

118 Statistical analyses of differences in soil solution concentrations were made using process mixed in 119 SAS 9.2. The 'Repeated' statement was used to account for the temporal dependency of successive 120 measurements close in time. DOC and DN concentrations were log-transformed before statistical 121 analysis. Back-transformed mean was estimated as $exp(mean + 0.5 s^2)$, with mean and variance, (s^2) , 122 on the log-transformed scale. Effects of tree species, block, sampling occasion and the interaction 123 between sampling occasion and treatment were included in the analyses.

124

125 To be able to estimate the total transport of C and N during the observed time period, together with a 126 confidence interval, we modeled concentrations for all series with the same tree species by a smooth 127 curve using a smoothing spline (Hastie et al. 2001). The smoothness of the curve was determined by 128 generalized cross validation (GCV). Residuals were computed as the difference between observed 129 concentrations and the smoothed curve and a large number of new series of concentrations were 130 created by randomly resampling the residuals and adding them to the original smoothed curve. For 131 each of these new series the concentrations were multiplied to the observed water discharge and 132 summed over all observations during the time period of the study. From these estimations of total transport of C or N, we collect the 2.5th and 97.5th percentile to define a 95% confidence interval for 133 134 the total transport. This is also called the bootstrapped percentile interval (Efron and Tibshirani, 135 1993). Fluxes are reported as annual means.

136

137 **3. Results**

138 **3.1 DOC**

139 The temporal variations in DOC concentrations in the O, A and B horizons are illustrated in Figure 1.

140 DOC concentrations under the O horizon in the birch stands were significantly lower than in the 141 spruce or pine stands (p=0.001 and 0.004, respectively) (Table 1). Average DOC concentration under

142 the O horizon in the birch stands was 25 mg L^{-1} and in the spruce and pine stands 43 and 39 mg L^{-1}

143 (Table 1). For all horizons, water fluxes were largest under birch and smallest under spruce and

144 therefore the relative difference between tree species was smaller for fluxes than for concentrations in

145 the O horizon. Annual fluxes were 38 g m⁻² y⁻¹ and 37 g m⁻² y⁻¹ for spruce and pine respectively and 21 g

146 $m^{-2} y^{-1}$ for birch (Figure 2).

147

148 Under the A horizon, there were no statistically significant differences between treatments for DOC 149 concentrations (Figure 1). Mean DOC concentration in the pine stands was 34 mg L⁻¹ and in the birch 150 and spruce stands 26 and 21 mg L⁻¹, respectively (Table 1). DOC fluxes were 27, 23 and 13 g m⁻² y⁻¹ for 151 pine, birch and spruce (Figure 2). The difference in DOC fluxes under the A horizon between spruce on 152 one hand and pine or birch on the other was statistically significant (p<0.05). The smallest water flux 153 occurred in the spruce stands, which explains the larger relative difference between fluxes than 154 between concentrations. DOC fluxes and concentrations thus decreased between the O horizon and the 155 A horizon in the coniferous stands – especially for spruce – but not in the birch stands.

156

In the B horizon the DOC concentrations were substantially lower than in the upper soil horizons for all tree species (Figure 1). Average DOC concentrations were 6, 5 and 8 mg L⁻¹ for birch, spruce and pine, respectively (Table 1). The corresponding fluxes were 4, 3 and 5 g m⁻² y⁻¹ (Figure 2). The difference between spruce and pine was statistically significant.

161

162 **3.2. DN**

163 DN concentrations in leachates under the O horizon were highest in the spruce stands, with on average 164 4,7 mg L⁻¹, whereas the DN concentrations in the birch and pine stands were 2.8 and 2.5 mg L⁻¹ 165 respectively (Table 1). This corresponds to fluxes of 3.5 g m⁻² y⁻¹ from the spruce stands and 2.4 and 2.2 166 g m $^{-2}$ y $^{-1}$ from birch and pine (Figure 2). The higher concentrations – but not the fluxes –in the spruce 167 stands, under the O horizon, were statistically significant at p<0.05 level. Under the A horizon, there 168 were no major differences between treatments. DN concentrations in this horizon ranged between 1.5 169 and 1.8 mg L⁻¹ (Table 1) and fluxes ranged between 1.2 and 1.4 g m⁻² y⁻¹ (Figure 2). In the B horizon the 170 DN concentration and flux in the spruce stands was significantly lower than for the other two species. 171 The DN concentrations under the B horizon were 1.9, 0.3 and 0.9 mg L-1 for birch, spruce and pine 172 respectively (Table 1). Fluxes were 0.9 g m⁻² y⁻¹ in the birch stands, 0.2 g m⁻² y⁻¹ in spruce stands and 173 $0.7 \text{ g m}^{-2} \text{ y}^{-1}$ in pine stands (Figure 2). 174

175 **3.3. pH**

- 176 The pH was significantly higher in the birch stands than in the coniferous stands under the O and A
- 177 horizons (Table 1). Mean pH in the birch stands were 5.0 and 4.7 in the O and A horizons, respectively
- 178 and in the spruce and pine stands 4.3-4.4 in both horizons. The same trend appeared also in the B
- horizon with lower pH under spruce (4.6) than under birch (4.8) and pine (4.7) (Table 1).
- 180

181 **3.4. SUVA**

SUVA₂₆₀ decreased significantly with DOC concentration and soil depth (Figure 3). Low DOC concentrations were associated with low UV absorbance, both compared across soil horizon and within each horizon. There were no major differences in specific UV absorbance between tree species at any depth (Table 1).

186

187 **4. Discussion**

188 The data presented here demonstrate that there were differences in DOC concentrations (Table 1) and 189 fluxes (Figure 2) under the O horizon between birch stands on one hand and pine and spruce stands 190 on the other. This difference was in line with expectations and the lower DOC concentrations and 191 fluxes under the O horizon in the birch stands are likely related to the thinner O horizons in these 192 stands. A previous study in the same stands showed that soil chemistry differs significantly between 193 tree species, as there were higher pH and base saturation in the birch stands (Hansson et al. 2011). In 194 addition that study showed that the O horizons in the birch stands were thinner than in the coniferous 195 stands and mixed with mineral soil. The effects of tree species on several chemical, physical and 196 biological properties of the soil is documented previously (e.g. Augosto et al. 2002; Binkley and 197 Giardina 1998). In other experiments with different tree species it has been shown that spruce forests 198 in Scandinavia have low pH and large C stocks in the O horizon (Oostra et al 2006; Vesterdal et al. 199 2008). The thinner O horizon in birch stands, both at our study site and in general, is likely related to 200 differences in litter composition between birch leaves and pine and spruce needles, resulting in more 201 bioturbation in the birch stands. This effect has been shown earlier by Saetre (1998), who reported that 202 mixing birch in spruce stands substantially increase earthworm activity.

203

204 The results may be compared to results from other studies. In a review of DOC in temperate forests, 205 Michalzik et al. 2001 did not find any difference in DOC and DON concentrations and fluxes in forest 206 floor leachates when comparing coniferous and hardwood sites. In an experimental forest in 207 Massachusetts, DOC was lower and DON and DIN fluxes higher from the O horizon under pine than in 208 the deciduous stands, but there were only small differences in DOC under the rooting zone (Currie et 209 al. 1996; McDowell et al. 2004). Other direct comparisons of tree species effects on DOC are scarce, 210 especially when interested in spruce, pine and birch, more specifically. There are data from Finland 211 comparing water extracts from the three species, indicating that there is less water soluble carbon 212 leached from pine stands than from spruce and birch (Smolander and Kitunen, 2002; Suominen et al., 213 2003). In a recent study, Lindroos et al. (2011) measured highest DOC concentration from pine stands 214 under the O horizon; significantly higher than under birch, with spruce having intermediate 215 concentrations.

216

217 DOC fluxes are perhaps more interesting to compare than concentrations. Water fluxes in the spruce 218 stands were substantially lower than in the birch and pine stands. Estimated annual water fluxes 219 under the O horizon were 940, 710 and 890 mm for birch, spruce and pine, respectively. The relative 220 difference increased with depth and under the B horizon estimated fluxes were 810 mm for birch, 550 221 mm for spruce and 690 mm for pine. This difference in water flux is related to differences in leaf area, 222 which has a large influence on transpiration and interception loss and was greatest for spruce. Low 223 water flux under spruce stands is also in accordance with other Scandinavian studies (Christiansen et 224 al. 2006, 2010). The relative difference in DOC flux from the O horizon between tree species was thus 225 lower than the relative difference in concentrations from the same horizon, but there were still larger 226 fluxes from the coniferous stands compared to the birch stands.

227

228 Contrary to our expectations, there was no difference in DOM quality as revealed by UV absorbance 229 between different tree species in any of the horizons (Figure 3), despite clear effects on soil chemistry 230 between the tree species (Hansson et al. 2011). DOC leached from litter from different tree species has 231 previously been shown to differ in SUVA (Hongve, 2000) and in water extracts from O horizons, birch 232 has been found to have higher concentration of phenolic compounds than spruce (Smolander et al. 233 2005). We had therefore hypothesized that there would be differences in DOC quality between the tree 234 species and especially between birch on one hand and the coniferous trees on the other. This was 235 however not the case, although this should not have been totally unexpected; other studies on water 236 extracts have shown that there are no or only minor differences in size and chemical fractions in 237 dissolved organic matter from birch, pine and spruce stands (Kiikkilä et al. 2006, 2011; Smolander et 238 al. 2005; Smolander and Kitunen 2002). One explanation for the small differences in DOC 239 composition between tree species quality could be that DOC largely originates in decomposed organic 240 matter in the lower parts of the O horizon (Fröberg et al. 2003, 2007b; Hagedorn et al. 2004), where 241 differences between birch leaves and coniferous needles are smaller than in fresh litter (Kiikkilä et al. 242 2011).

243

244 Whereas there were no effects of tree species on DOC quality, there was a positive relationship 245 between DOC concentration and SUVA and furthermore also a trend with decreasing SUVA with depth 246 in the soil profile. The same trend with SUVA changing with depth has been reported previously by 247 Hagedorn et al. (2000), Don et al. (2008) and Sanderman et al. (2008, 2009) who explained this by 248 additions of DOC with low SUVA at depth. An alternative explanation would be preferential adsorption 249 of hydrophobic organic matter, but this is likely not the case as isotope data have shown that DOC 250 from fresh litter is retained in the upper part of the soil and replaced by old carbon from the mineral 251 soil (Fröberg et al. 2007a, Sanderman et al. 2008, Tate et al. 2011). This strong physico-chemical 252 control of DOC by minerals also explains why the difference in DOC concentrations and fluxes under 253 the O horizon between tree species was not found in the mineral soil in our study.

254

Increasing pH as an explanation of rising DOC concentrations in streams and lakes in large areas inEurope and North America has got much attention during the last decade (e.g. Monteith et al. 2007;

257 Roulet and Moore, 2006). According to the hypothesis, decreasing acid inputs have resulted in 258 increasing pH and as a consequence increasing DOC concentrations in surface waters (Evans et al. 259 2006). The effect is thus opposite to what we found here, i.e. lower DOC concentration and fluxes from 260 the O horizon of the birch stands where pH was significantly higher than in the coniferous stands. 261 Although it is well established that organic matter solubility increases with increasing pH, this effect 262 was not manifested in our study. As discussed above, we instead suggest that the differences in soil 263 chemistry under different tree species, in particular the higher pH under birch, was one reason for the 264 thinner O horizon in the birch stands under this horizon. This in turn is likely to result in lower DOC 265 concentrations (Fröberg et al. 2005).

266

267 The lower DOM flux from the O horizon under birch than in coniferous stands, in combination with 268 similar DOM flux in the mineral soil regardless of dominant tree species, imply that there is a lower 269 transport of DOM from the O horizon to the mineral soil in the birch stands. This could have 270 implications for long-term C budgets in the mineral soil. As reported previously (Hansson et al 2011) 271 there was a significant difference in O horizon SOC stocks between the tree species, with largest C 272 stock in the spruce stands and lowest C stocks in the birch stands. The differences in the mineral soil 273 were however not as obvious. As the C stocks in the mineral soil change only slowly, an effect on the C 274 stocks was not to expect. However, if extrapolated into the future a difference in DOC influx to the 275 mineral soil could have implications also for mineral soil C stocks. Biological mixing is however 276 probably of greater relative importance in the birch stand than in the coniferous stands and it thus 277 seems like in pine and spruce forests, DOC is a more important pathway for translocation of carbon 278 from the upper to the lower parts of the soil profile when compared to birch forests, whereas the 279 opposite may be true for physical mixing as a means of transporting soil carbon from the surface to 280 deeper horizons.

281

282 The DN data show that a considerable amount of N was leached through the soil. Furthermore, low 283 C:N ratios in solution coincide with high DN, which shows that a significant fraction of the N was 284 transported in inorganic form at all sampled depths, i.e. all the way down below the B horizon (Figure 285 4). This was also obvious from a few measurements (four sampling occasions) of ammonium and 286 nitrate (data not presented). Effects of tree species on N leaching has been studied previously and one 287 typical observation is greater influx of N via deposition in coniferous species compared to deciduous, 288 but no major differences between coniferous and deciduous stands in N leaching from the soil 289 (Christiansen et al 2010; Gundersen et al. 2009). This is in accordance with the data presented here. 290 The nitrogen flux under the O horizon was highest under spruce, but more N was leaving the rooting 291 zone from the birch and pine stands than from the spruce stands (Figure 2). Accordingly, more N is 292 circulating in the spruce stands than in the pine or birch stands. The spruce stands seem to have the 293 ability to retain N within the system, whereas N is apparently lost from the pine and birch stands 294 (Figure 4). In a previous study, it was demonstrated that the total soil organic N pool at our study site 295 was greatest under in the spruce stands (Hansson et al. 2011). Approximately 100 g more N per square 296 meter had accumulated in soil organic matter in the spruce stands than in the birch stands, already 297 approximately 50 years after the stands were established. This difference in N stock corresponds to 2 g

- 298 N m⁻² for every year since the establishment of the experiment. We do not have the ambition, in this
- study, to make a complete N budget for the different stands. However, the difference in soil N stocks
- 300 presented by Hanson et al. (2011), in combination with the data presented here, suggest a strong link
- 301 between tree species, soil organic matter accumulation and N leaching. The flux of dissolved N in
- 302 leachates from the birch and pine stands corresponds to about half of the annual N deposition or more,
- 303 whereas the amount of N leached from the spruce stands was minor, in accordance with the large
- 304 differences in soil N pools at this site, previously presented by Hansson et al. (2011).
- 305

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Tables

Table 1. Concentrations of DOC and DN (mg L^{-1}), specific UV absorption (SUVA) at 260 nm (L mg⁻¹ m⁻¹) and pH.

	Horizon		Birch	Spruce	Pine
DOC	0	Mean	25 (20-31) a	43 (36-51) b	39 (33-46) b
	А	Mean	26 (17-42) a	21 (14-29) a	34 (24-49) a
	В	Mean	6 (4-8) ab	5 (4-6) a	8 (6-10) b
DN	0	Mean	2.8 (2.3-3.5) a	4.7 (3.9-5.6) b	2.5 (2.1-3.0) a
	А	Mean	1.8 (1.2-2.5) a	1.6 (1.2-2.2) a	1.5 (1.1-2.1) a
	В	Mean	1.9 (1.1-3.0) a	0.3 (0.2-0.5) b	0.9 (0.6-1.3)a
SUVA	0	Mean	4.6 (4.2-5.1) a	4.5 (4.2-4.9) a	4.5 (4.1-4.9) a
	А	Mean	3.9 (3.5-4.4) a	3.4 (3.0-3.8) a	3.8 (3.4-4.1) a
	В	Mean	2.2 (1.7-2.6) a	2.1 (1.7-2.4) a	2.6 (2.3-3.0) a
рН	0	Mean	5.0 (4.9-5.2) a	4.4 (4.2-4.5) b	4.3 (4.1-4.5) b
	А	Mean	4.7 (4.5-4.9) a	4.4 (4.2-4.5) b	4.3 (4.2-4.5) b
	В	Mean	4.8 (4-7-4.9) a	4.6 (4.5-4.7) b	4.7 (4.6-4.8) a

Figure captions

Figure 1. DOC concentrations in soil water collected with lysimeters under the O, A and B horizon. Error bars represent standard deviations (n=2 or 3).

Figure 2. Fluxes of dissolved organic carbon (DOC) and dissolved nitrogen (DN). Error bars represent the 95% confidence interval (95% bootstrapped percentile interval).

Figure 3. Specific UV absorbance (SUVA) as a function of DOC concentration. Filled black symbols represent the O horizon, grey symbols the A horizon and open symbols the B horizon.

Figure 4. Total dissolved nitrogen (DN) and Dissolved organic carbon/DN ratios (C/N) under the B horizon for birch, spruce and pine. High DN concentrations coincide with low C/N ratios, strongly suggesting a large proportion of inorganic N.







Figure 2.







Figure 4.