Page 1 of 59 Oikos

1	Title : Ecological strategies in stable and disturbed environments depend on species
2	specialisation
3	
4	Authors : Lucie Büchi ^{1,*} and Séverine Vuilleumier ^{1,2}
5	¹ Department of Ecology and Evolution, and Institute of Microbiology
6	University of Lausanne
7	1015 Lausanne - Switzerland
8	² School of Life Sciences,
9	Ecole Polytechnique Fédérale de Lausanne,
10	1015 Lausanne, Switzerland
11	
12	*Corresponding author, current address: Agroscope, Institute for Plant Production Sciences
13	1260 Nyon, Switzerland
14	
15	Lucie.Buchi@gmail.com
16	Severine.Vuilleumier@epfl.ch
17	
18	
19	Keywords: Specialists, Generalists, Metacommunity, Competition-reproduction trade-off,
20	Dispersal, Generation overlap

Abstract

Ecological strategies are integral to understanding species survival in different environments. However, how habitat specialisation is involved in such strategies is not fully understood, particularly in heterogeneous and disturbed environments. Here, we studied the trait associations between specialisation, dispersal ability, competitiveness, reproductive investment and survival rate in a spatially explicit metacommunity model under various disturbance rates. Though no unique trait values were associated with specialisation, relationships were uncovered depending on environmental factors. We found strong trait associations mainly for generalist species, while specialist species exhibited a larger range of trait combinations. Trait associations were driven first by the disturbance rate and second by species' dispersal ability and generation overlap. With disturbance, low dispersal ability was strongly selected against, for specialists as well as for generalists. Our results demonstrate that habitat specialisation is critical for the emergence of trait strategies in metacommunities and that disturbance in interaction with dispersal ability limits not only the range of trait values but also the type of possible trait associations.

Page 3 of 59 Oikos

Introduction

37

38 Characteristics such as size, dispersal ability, reproduction, or survival rate, vary drastically 39 among species. For example, the magnitude of divergence of seed size and mass among flowering plants can reach 10¹¹ and 10⁵, respectively (Westoby et al. 1992). Similarly, 40 41 flowering plants' longevity ranges from desert annuals, completing their life cycle in a few 42 weeks, to thousand-year-old trees such as the bristlecone pine *Pinus longaeva* (Borges 2009). 43 These variations in life history or functional trait values are shaped by the evolutionary 44 history of the species; for a given species, the association of particular trait values defines its 45 ecological strategy, or syndrome (Westoby 1998, Reich et al. 2003, Grime and Pierce 2012). 46 A strategy is the result of selection for trait values whose combination confers a higher fitness 47 in a given environment (Agrawal et al. 2010, Stevens et al. 2014). This selection can act on 48 independent traits, or on traits constrained by trade-offs, narrowing the range of possible trait 49 combinations. 50 Ecological strategies have been thoroughly studied and several frameworks focusing therein 51 have been developed, from the classical r/Kselection theory (MacArthur and Wilson 1967, 52 Pianka 1970) and Grime CSR classification for vascular plants (Grime 1977), to the more 53 recent leaf-height-seed strategy scheme and leaf economics spectrum (Westoby 1998, 54 Westoby et al. 2002, Wright et al. 2004, Adler et al. 2014, Reich 2014). The description of 55 species' ecological strategies is intimately linked to their environment, especially in terms of the intensity of stress and disturbance experienced (Grime 1977, Southwood 1988, Grime and 56 57 Pierce 2012, Westoby 1998). 58 Habitat specialisation also varies among species: at one extreme, specialists exhibit a narrow 59 niche breadth and/or environmental tolerance while generalists have a larger breadth and/or 60 tolerance (Levins 1968, Futuyma and Moreno 1988, Poisot et al. 2011). This degree of 61 specialisation changes how species experience their environments; for example, a given

62	environment may be perceived as more homogeneous by a generalist species than by a
63	specialist one. As a consequence, specialisation may also influence the selection on trait
64	values underlying ecological strategies. Despite the likelihood of such influence, few studies
65	have investigated these links specifically. Most studies have focused on the relationship
66	between specialization and dispersal. Empirically, specialists generally have a low dispersal
67	rate (Fig. 1A) (Warren et al. 2001, Verberk et al. 2010). Large scale empirical studies in
68	European land snails (Dahirel et al. 2015) and in European birds (Reif et al. 2015) also tended
69	to associate specialist species with low dispersal. Despite this, specialisation has
70	paradoxically been associated with high dispersal abilities (Fig. 1B), particularly when
71	specialisation concerns sparse or ephemeral habitats (Levin and Muller-Landau 2000, Fridley
72	et al. 2007). Stevens et al. (2014) have also suggested that while a link between specialisation
73	and dispersal is expected, it is difficult to predict the direction of this correlation.
74	Theoretically, studies have generally shown that generalist species are commonly associated
75	with disturbed and heterogeneous environments, and specialist species with stable and
76	homogeneous habitats (Futuyma and Moreno 1988, Devictor et al. 2008). In parallel,
77	investigations about the coevolution of life history traits have suggested that low rates of
78	dispersal are required for specialisation to evolve (Brown and Pavlovic 1992, Kisdi 2002,
79	Ravigné et al. 2009, Débarre and Gandon 2010, Nurmi and Parvinen 2011). Other studies
80	have suggested that specialists could be favoured at intermediate dispersal rates (Fig. 1C)
81	(Ronce and Kirkpatrick 2001) or, inversely, that they can be advantaged by small or large
82	values (Fig. 1D) (Nurmi and Parvinen 2008). However, analysing the conditions required for
83	specialisation to evolve is not equivalent to investigating the traits leading to the most
84	efficient strategies, especially in changing environments. The lack of consensus on the
85	direction of covariation between dispersal ability and specialisation may well be due to the
86	multiple roles of dispersal seen in life-history strategies (Poisot et al. 2011, Stevens et al.

88

89

90

91

92

93

94

95

96

97

98

99

100

101

102

103

104

105

106

107

108

109

110

111

2014), as well as different dispersal attributes (rate, distance and temporality) and phases (Bonte et al. 2012, Buoro et al. 2014). Less evidence has been shown regarding the links between specialisation and other life history traits. Reif et al. (2015) have revealed complex relationships between habitat specialisation, diet and climatic niches and flight attributes. Specialists are also traditionally associated with K-strategy characteristics (Southwood 1988), such as high competitive ability (Fig. 1EF) (Marvier et al. 2004). Specialists may also have longer life spans than generalists (Fridley et al. 2007). These examples demonstrate that there is little consensus on the association between the degree of specialisation and other traits, and that these aspects deserve further investigations. Most of the theoretical studies investigating the coevolution of life history traits and ecological strategies include few habitat types and few competing species, mainly due to difficult analytical tractability. Moreover, important life history traits other than dispersal, such as survival rate or competitiveness have seldom been considered. Understanding the evolution of specialisation and related strategies therefore deserves investigation in wider settings. A simulation approach considers the multiplicity and complexity of trait interactions involved in building ecological strategies. In this study, we use simulation experiments within spatially explicit environments to investigate the selection of strategies in metacommunities under different disturbance regimes. The modelling approach mimics community assembly processes. Each simulation starts with the creation of a large pool of species harbouring different strategies, which are then selected by environmental and spatial processes. Strategies combine different values of habitat specialisation and dispersal ability, competitiveness and reproductive investment. We analyse the resulting trait value associations across different values of generation overlap and disturbance. We expect that different degrees of specialisation are associated with distinct strategies. In particular, we investigate the

relationship between specialisation and dispersal and whether generalist species show trait attributes linked to opportunistic strategies such as low competitiveness (Figure 1). We predict, by modifying the intensity of competition and the spatial distribution of available habitats, that both disturbance rate and generation overlap impact the trait associations observed.

117

118

119

120

121

122

123

124

125

126

127

128

129

130

131

132

133

134

135

112

113

114

115

116

Methods

Model description

We used a metacommunity simulation model in which trophically-equivalent species (e.g. all plants) compete for space in a heterogeneous environment (Büchi et al. 2009, Büchi and Vuilleumier 2014). This model simulates sessile species, in which only juveniles dispersed prior to settlement. The environment is composed of 25 x 25 local communities, or cells, each cell being characterized by a carrying capacity K of 100 individuals and an environmental value E_i that determines species growth rate. E_i varies from cell to cell (heterogeneous environment) and follows a normalized gaussian probability distribution (with mean = 0 and standard deviation = 1, Büchi et al. 2009). Metacommunity dynamics proceed in four discrete time steps: 1. reproduction, 2. adult mortality and disturbance, 3. juvenile dispersal and 4. competition for space. Each species is characterized by six traits: niche optimum μ_s , niche breadth σ_s , dispersal ability δ_s , competitiveness κ_s , reproductive investment ω_s and survival rate ψ_s . Species' habitat specialization impacts their fecundity, with specialist species having a higher fecundity in their optimal habitat compared to generalist species. Species fecundity R_s in a community (cell) of environmental value E_i is determined by the individual reproductive investment ω_s , niche optimum μ_s , and niche breadth σ_s as follows:

Page 7 of 59 Oikos

136
$$R_s(E_i) = h \cdot \frac{\omega_s}{\sigma_s \sqrt{2\pi}} \cdot \exp\left[-\frac{1}{2} \cdot \left(\frac{E_i - \mu_s}{\sigma_s}\right)^2\right]$$
 (1)

- Here h is a scaling factor transforming the rate of increase into an effective fecundity (h is set
- to 10 in this study). The effective per capita number of offspring in each cell is the rounded
- value of $R_s(E_i)$.
- Mortality occurs after reproduction and is determined by adult survival ψ_s and disturbance
- rate T. Adult survival probability equals to ψ_s at each time step. Disturbance causes total
- extinction in proportion *T* of randomly chosen local communities (cells).
- Juvenile dispersal occurs according to a dispersal kernel $D_s(x)$ (probability density function of
- dispersal distance) whose shape is defined by the mean species dispersal ability δ_s ,

145
$$D_s(x) = \frac{1}{\delta_s} \cdot \exp\left[-\frac{x}{\delta_s}\right]$$
 (2)

- Each juvenile disperses independently at a distance that is determined by the dispersal kernel
- associated with its species. Direction is drawn randomly from a uniform distribution. Periodic
- boundary conditions are considered, such that individuals reaching the environment borders
- re-enter the environments from the opposite side.
- Preemptive competition is assumed (i.e., adult individuals already settled cannot be displaced
- by juveniles). Therefore, only juveniles compete for settlement in each local community;
- successful juveniles are drawn randomly among the pool of competing juveniles, with a
- weighting determined by their competitiveness κ_s . If there are fewer competing juveniles than
- the number of places available, all juveniles can settle.
- We further assume that competitiveness κ_s and reproductive investment ω_s are constrained by
- the following trade-off (Levine and Rees 2002):

 $157 K_s \cdot \omega_s = 1 (3)$

This trade-off is inspired by the classic, well-documented trade-off between seed size and seed number (Jakobsson and Eriksson 2000, Leishman 2001).

In this model, traits are randomly associated in species and advantageous strategies are selected when providing a fitness advantage to species. During this process, trait values remain constant. We thus assume that the time scale of the processes studied does not allow for the evolution of traits, and only community dynamics are considered.

This model is an object-oriented model implemented in Borland Delphi, and code is available upon request.

Simulations

For each simulation, we generated an initial metacommunity composed of species to which we assigned trait values as described in Table 1. Three distinct scenarios were built in order to address the influence of the different species traits sequentially. In the first scenario, all species had competitiveness κ_s and reproductive investment ω_s set to 1 and had variable dispersal ability δ_s drawn randomly between 0 and 1. In the second scenario, all species had variable competitiveness κ_s and reproductive investment ω_s , which were randomly drawn for each species following the trade-off presented in equation 3 (uniform distribution of their logarithms), and had an identical and limited dispersal ability δ_s set to 0.1. In the third scenario, all species had variable dispersal ability, competitiveness and reproductive investment. In all scenarios, we varied for each species niche optimum μ_s (values between - 2.5 and 2.5) and niche breadth σ_s (values between 0.01 and 1). A value of 0.01 for σ_s corresponded to very specialist species and a value of 1 corresponded to very generalist species. Then, for each scenario, we investigated the influence of generation overlap (adult annual survival) and disturbance as follows. Each scenario was run alternatively with adult

Page 9 of 59 Oikos

182	annual survival rate ψ_s set, to either 0 (no generation overlap) or to 0.95 (strong generation
183	overlap). Three rates of disturbance T were considered: $T = 0$, $T = 0.01$ and $T = 0.25$.
184	For each case studied, we considered pools of 100 different species to initiate the simulations.
185	Individuals of these species were randomly distributed in the landscape until carrying capacity
186	was reached in each cell. Metacommunity dynamics were simulated for 1000 time steps,
187	during which some species went extinct and others persisted. This duration is sufficient to
188	reach a steady state in terms of mean traits (weighted by species abundances) though the
189	species richness continued to decrease very slowly (Supplementary material Appendix 1). For
190	each of the simulation scenarios, 50 replicates were run, with a new species pool and
191	landscape for each replicate. At the end of each simulation, we recorded the characteristics
192	and abundance of the selected strategies in terms of niche breath, dispersal ability,
193	competitiveness and reproductive investment. Mean trait values were computed for each
194	simulation considering the trait values of all the surviving individuals (mean species trait
195	value weighted by the abundance of each species). Results for trait associations given niche
196	breadth were analysed for five categories of niche breadth (from the most specialist to the
197	most generalist species): 0-0.1, 0.1-0.2, 0.2-0.4, 0.4-0.6, 0.6-1.
198	We also investigated the sensitivity of our results to the initial number of species (10 or 1000)
199	and to different environmental spatial autocorrelation. To generate autocorrelation between
200	the environmental values of two cells, we used the sequential Gaussian simulation algorithm
201	(Goovaerts 1998, see also figure 1 in Büchi et al. 2009). We considered moderate ($\alpha = 5$) and
202	strong (α = 10) autocorrelation, where α represents the distance above which correlation falls
203	below 0.05 (α = 0 in the other simulations). These results are presented in Supplementary
204	material Appendix 2 and 3.
205	Simulations results were analysed using R 3.0.2 (R Core Team 2013).

Resu	lts
------	-----

Associations between niche breadth and dispersal when competitiveness and reproductive
investment are fixed
Metacommunities are mainly composed of species with low values of niche breadth, namely
specialist species (Fig. 2). Interestingly, a strong association between niche breadth and
dispersal abilities is observed (Fig. 2). In the absence of disturbance ($T = 0$), communities are
composed of a high number of specialist species with a large range of dispersal abilities, but
with very low abundances at the lowest dispersal values, while more generalist species have
lower dispersal abilities (Fig. 2). The presence of disturbance affects the association between
species niche breadth and dispersal abilities, such that specialist and generalist species with
limited dispersal values are progressively driven to extinction (Fig. 2). Generation overlap
changes this pattern only quantitatively, except at high disturbance rate where the pattern is
completely changed. In this situation, mean dispersal is higher than in any other situation, and
generalist species have higher dispersal abilities than specialist ones (Fig. 2).
The associations found between niche breadth and dispersal are quantitatively but not
qualitatively impacted by spatial autocorrelation (Supplementary Material Appendix 3). An
exception occurs when disturbance is high $(T = 0.25)$ and generations overlap. In this case,
when spatial autocorrelation is low, dispersal was lower for all species, though this effect was
stronger for the most generalist species.
In the absence of disturbance, species can survive with a large range of niche optimum values
whereas species with extreme niche optima (thus relying on rare habitats) disappear in the
presence of strong disturbance (Supplementary material Appendix 4).
Associations between niche breadth and competitiveness/reproductive investment when

dispersal ability is limited

233

234

235

236

237

238

239

240

241

242

243

244

245

246

247

248

249

250

251

252

In the absence of disturbance (T=0), metacommunities are composed of species with a large range of competiveness and niche breadth, but with a higher proportion of species exhibiting small niche breadth values (Fig. 3). With no generation overlap, generalist species show reduced investment in competitiveness compared to specialist species with a large range of possible competitiveness and a mean value similar to the initial state (Fig. 3). When generations overlap, all categories of specialisation show both a mean and range of values similar to the initial state, with no clear selection of a particular strategy. Disturbance tends to decrease global investment in competitiveness in favour of reproduction. When disturbance is moderate (T = 0.1), the association between competitiveness and niche breadth is still negative, with competitiveness being globally lower with generation overlap. In contrast, at high disturbance rate (T = 0.25), all categories of niche breadth show similarly low investment in competition, with a reduced range of viable trait values (Fig. 3). This pattern is particularly strong when generations overlap. Here again, spatial autocorrelation has a slight quantitative influence on the relationship between niche breadth and competitiveness (Supplementary Material Appendix 3). However, when disturbance was high (T = 0.25), strong spatial autocorrelation allowed for a larger range of competitiveness values for specialist species. With limited dispersal ability, very generalist species can survive in the metacommunity. They are however, in the absence of disturbance, restricted to extreme niche optima. In contrast, very generalist species with intermediate niche optimum can survive in the presence of disturbance (Supplementary material Appendix 4).

253

254

Associations between niche breadth, dispersal and competitiveness/reproductive investment

The change from fixed values of competitiveness and reproductive investment to variable
values has almost no influence on the association between dispersal ability and niche breadth
(Fig. 4 vs Fig. 2).
In contrast, the change from fixed values of dispersal ability to variable values has a clear
influence on the association between competitiveness and niche breadth (Fig. 5 vs Fig. 3).
First of all, dispersal limitation allows for the survival of a larger number of generalist species
than a variable dispersal ability. Then the decrease of competitiveness with increasing niche
breadth is clearly less pronounced when dispersal is variable (Fig. 5). The range of
competitiveness values observed in the presence of disturbance is also clearly higher,
especially at low niche breadth. So here, under disturbance and depending on dispersal
limitation, communities can be either composed by specialist species with a broad range of
competitiveness/reproductive investment values or by generalist and specialist species that
invest in reproduction and limit their investment in competitiveness.
Associations between dispersal and competitiveness/reproductive investment are visible when
taking into account the niche breadths of the surviving species, especially at high disturbance
rates (Fig. 6). When none or moderate disturbance occurs and there is no generation overlap,
generalists associate higher reproductive investment with lower dispersal ability compared to
specialists. When no disturbance occurs and generation overlap is high, generalists have lower
dispersal ability but with no clear selection on competitiveness values. The pattern changes
when disturbance rate is strong and generation overlap is high. In this case, generalist species
have higher dispersal ability with higher reproductive investment than specialist species (Fig.
6).

Discussion

Strategies involving specialisation and dispersal

Strategies involving specialisation and dispersal abilities are dependent on the disturbance
rate. In the absence of disturbance, our results show that generalist species have limited
dispersal ability, while in contrast, specialist species are associated with a large range of
dispersal abilities. These results can be explained first by the cost of dispersal and second by
the amount of suitable habitat accessible. In the absence of disturbance, dispersal is costly
because the probability of a juvenile reaching an unsuitable habitat is high (Bonte et al. 2012),
and thus limited dispersal abilities develop (Hastings 1983). This result is observed here for
generalist species but not for specialist species. Indeed, specialist species produce more
juveniles than the number of local settlement opportunities; dispersal therefore favours their
persistence despite its inherent cost by allowing them to reach new habitats. In addition
limited dispersal ability drastically reduces specialist species' chances to settle in their
specific, rare and scattered suitable habitat. This result is reflected in the lower abundance
observed for the most specialist species at very low dispersal abilities. In contrast, generalist
species are less impacted by the distribution of their habitats, which tend to be widely
distributed and could be easily reached through dispersal from neighbouring habitats (Nilsson
et al. 1993, Dynesius and Jansson 2000).
The presence of disturbance strongly modifies the association between species niche breadth
and dispersal abilities. Under disturbance, specialist and generalist species with limited
dispersal abilities are excluded, as disturbance induces extinction-colonization dynamics
favouring species with high dispersal (Gandon and Michalakis 1999). Generalists face
additional barriers as well, such as their very low growth rate, which does not withstand
frequent disturbance (Parvinen and Egas 2004, Nurmi and Parvinen 2008). The coexistence of
specialist and generalist species is also driven by generation overlap. When adult survival is
null, recruitment to suitable habitat is high and specialist species dominate over generalist
species as they produce more juveniles. In contrast, when generations strongly overlap,

saturation of the metacommunity is very high. In such a saturated system, disturbance creates
gaps in adult density, and thus many spots are available locally for juveniles. In this case,
species with high dispersal ability have a strong advantage in finding these newly opened
positions. The production of numerous offspring can confer similar advantages, explaining
why very fecund specialists can perform well even when they have lower dispersal ability.
Some examples are tropical forests or coral reefs, where sudden gaps are often the only way
for new recruitment to occur (Connell 1978).
These results contrast the classical descriptions of specialists as having reduced dispersal
ability and generalists as being opportunistic species with high dispersal ability. This classic
view is supported by theoretical studies on the evolution of specialisation, or local adaptation,
in interaction with dispersal (Brown and Pavlovic 1992, Kisdi 2002, Ravigné et al. 2009).
Those studies demonstrate that reduced dispersal is required for specialisation, as dispersal
prevents adaptation to a local environment. As a corollary, high dispersal promotes generalist
species by increasing the variety of habitats a species experiences. The discrepancy between
these classic descriptions and the results seen here can be explained by differences in the
processes studied. Here, we investigated the dynamics and factors that drive specialist-
generalist coexistence in metacommunities and not how specialisation itself might evolve
(e.g. from an initial pool of generalist species). Although dispersal limitation favours
specialists, there is still a need to disperse in order to persist in dynamic and heterogeneous
environments. Interestingly, the diversity of trait associations observed in our study is
reflected by some empirical evidence. Both generalist and specialist species can have variable
dispersal abilities (e.g. Nilsson et al. 1993, Levin and Muller-Landau 2000, Warren et al.
2001, Verberk et al. 2010).

Strategies involving specialisation and competitiveness/reproductive investment

Strategies involving specialisation and competitiveness (or reproductive investment) are less
obvious than those involving dispersal, and are dependent on the disturbance rate in
interaction with generation overlap and dispersal ability. In the absence of disturbance and
when generation overlap is strong, recruitment is low and the coexistence of specialist and
generalist species is independent of their investment in competition or reproduction. Forests
are a typical example of a natural ecosystem with high generation overlap, as they are mainly
composed of long-lived species. In contrast, without generation overlap, recruitment is high
and generalists investing in competition rather than reproduction are excluded by specialist
species (Nurmi and Parvinen 2008). Specialist species, which produce many juveniles in their
given habitat, can afford to invest more in competitive ability, while generalist species are
strongly limited by their reduced fecundity (Nurmi and Parvinen 2008).
Disturbance has a drastic impact on the associations between niche breadth and
competitiveness or reproductive investment, but the patterns depend strongly on the dispersal
ability of the species. When dispersal is unconstrained, communities are composed of
specialist species, with generalists excluded. Disturbance creates extinction-colonization
dynamics that are known to favour dispersal, and provides empty habitat patches that are
efficiently colonized by specialist species due to two factors: their local fitness advantage and
the larger number of dispersing juveniles. Thus, disturbance favours species with high
dispersal and fecundity as they efficiently settle in empty habitat patches. Dispersal limitation,
however, completely shifts this pattern: though new habitat becomes available after
disturbance, the sparseness of suitable habitat for specialists and their limited ability to
disperse prevents occupation of these new areas, and thus prevents displacement of the (less
efficient) generalists. Communities are composed of species with a wide range of
specialisation but usually with all species investing in reproduction. Species with high
competitiveness go extinct due to their corresponding low colonization notential

379

380

381

382

383

384

385

386

387

388

389

390

391

392

393

394

395

396

397

398

399

400

401

402

Our results show thus that specialist and generalist species can either invest in reproduction or in competition, except in the condition of high disturbance rate with limited dispersal, where investment in reproduction is critical for survival. Our results therefore contrast with the common conclusion that specialist species benefit from "K-strategy" characteristics such as high competitive ability, while generalists are more associated with "r-strategy" characteristics such as low competitive ability (Southwood 1988). We also find that competition and reproductive investment play less of an important role in specialist-generalist coexistence and in the emergence of trait strategies, when compared to the fundamental role of dispersal. Strategies involving specialisation, dispersal and competitiveness or reproductive investment In agreement with the results obtained by Ronce (2000b), no global association was observed between dispersal and competitiveness/reproductive investment when dispersal varies along with competitiveness and reproductive investment. However, interestingly, an association between these traits emerges when species are grouped by specialisation levels. When no disturbance occurs and there is no generation overlap, recruitment is high and homogeneous, favouring species with both reduced dispersal and increased reproductive investment (Ronce and Olivieri 1997). In contrast, when disturbance rate and generation overlap both occur, generalist species disperse effectively and invest greatly in reproduction. This situation is similar to gap creation in rain forests or coral reefs, leading to a "fugitive species syndrome" (Connell 1978, Tilman 1994). Our results demonstrate that there is not a unique association between specialisation, dispersal ability and reproductive investment. These associations are instead driven by recruitment

opportunities that are determined by disturbance rate and generation overlap. Similar results

Oikos

were obtained by Venable and Brown (1988), who show that the relative importance of seed traits (e.g. size, dispersal, dormancy) depends on dispersal limitation, spatial and temporal autocorrelation and disturbance.

406

407

408

409

410

411

412

413

414

415

416

417

418

419

420

421

422

423

424

425

426

427

403

404

405

Model discussion and perspectives

The approach adopted in our model consisted of generating species rich metacommunities in which each species has random trait values. We subsequently observed the selection of strategies that allowed species survival. This corresponds to community assembly processes, which are adopted less often than genetic approaches (but see for example Kallimanis et al. 2006, Devictor and Robert 2009). During simulations, traits did not evolve and the invasion of new or formerly extinct species was not allowed. We assumed that the time scale of the processes studied did not allow for trait evolution, as, our system is meant to mimic, for example, the colonisation of a novel habitat by competing species or a community experiencing a rapid change of its environment, as expected with anthropogenic changes. The use of a simulation model was necessary given the multiplicity of the traits considered and number of processes involved, as such a system would not be analytically tractable. Simulations also allow a thorough characterization of the successful strategies under different conditions. Finally, our simulation framework complements common models used to test theoretical metacommunity paradigms (see review in Logue et al. 2011). Here, we considered 50 replicates of simulations, each starting with 100 species. This translated into a random sample of 100 x 50 combinations of trait values. This procedure allowed for the exploration of a large range of trait values. The number of species considered in our simulation set up is representative of what is generally observed in natural communities. Indeed, species richness is generally around 20-40 species and rarely goes beyond 100 species in temperate grasslands and forests (e.g. Proulx and Mazumder 1998,

428	Klimek et al. 2007, Morin et al. 2011, Axmanova et al. 2012). Accounting for a larger number
429	of initial species does not qualitatively change the results (see results for 1000 species in the
430	Supplementary Material Appendix 2). However, it has the consequence of generating many
431	highly specialised species that occupy the whole range of environmental space. This process
432	drives generalist species to extinction; we therefore cannot effectively describe strategies for
433	generalist species under this condition. When a reduced number of species is considered (see
434	results for 10 species in the Supplementary Material Appendix 2), the resulting
435	metacommunity poorly represents the potential diversity of strategies and combine few traits
436	values. In this case, competition is very low and almost all species are maintained in the
437	metacommunity, and strategies are thus not easy to identify.
438	In our simulations, the associations found between niche breadth and other traits are generally
439	quantitatively but not qualitatively impacted by spatial autocorrelation (see Supplementary
440	Material Appendix 3). The main differences are observed in the presence of disturbance. We
441	expect a more drastic effect if this environmental autocorrelation is associated with
442	autocorrelation in disturbance (as predicted for fire, hurricane, or drought). Indeed it has been
443	shown, in a metapopulation context, that the effect of autocorrelated disturbance can interact
444	with habitat configuration (Vuilleumier et al. 2007). In this situation, specialist species might
445	be driven to extinction by aggregated disturbance in their clustered habitats while generalist
446	species might find refuge in more type of habitats, allowing re-colonisation following
447	disturbance events.
448	The exact form of a trade-off can influence the species coexistence outcome (Levins 1968, Yu
449	and Wilson 2001, Levine and Rees 2002, Nurmi and Parvinen 2008). Our simulations
450	considered two trade-offs: one with habitat specialisation, and the other with investment in
451	competition and reproduction. The specialisation trade-off considered here is based on the
452	species growth rate, which might be more detrimental to generalist species than specialist

453	ones (Nurmi and Parvinen 2008). The trade-off between competition and reproduction
454	modelled here was inspired by the well-documented trade-off between number and size of
455	flowering plants' seed (Leishman 2001). While this trade-off has been shown to promote
456	coexistence (Tilman 1994, Yu and Wilson 2001), its role in shaping natural communities'
457	composition, and its consequences in terms of species performance, have often been
458	questioned (Leishman 2001, Moles and Westoby 2006, Agrawal et al. 2010). Moreover,
459	trade-offs are expected to vary among species and depend on the environments experienced
460	by the species (Agrawal et al. 2010). How such variability impacts species' coexistence in a
461	community remain to be investigated.
462	Our results show that dispersal is a critical factor in shaping trait associations in
463	metacommunities. To gain further insights on its impact on community composition, we
464	suggest that dispersal rate and dispersal distance should be considered as two independent
465	factors. We expect that specialist species would benefit from rare, long distance dispersal
466	events but do not require a high dispersal rate.
467	Specialist species are currently declining worldwide (Warren et al. 2001, Munday 2004,
468	Clavel et al. 2010). However, our results demonstrate that specialists could be successful even
469	in the presence of disturbance. This counterintuitive result could be explained by the large
470	availability of habitats for specialist species in our study, while in natural ecosystems,
471	specialist species suffer from important habitat loss. The impact of habitat loss on the
472	coexistence of specialist and generalist species, and on the association of specialisation with
473	the other traits, especially dispersal ability, would thus deserve further investigation.
474	Very few experimental investigations on the association of specialisation and other life
475	history traits exist. However, recently, two experimental studies on trait co-variation with
476	specialisation reported interesting and unexpected results. First, Khokhlova et al. (2014)
477	demonstrated, using fleas, that stronger trade-offs between quantity and quality of offspring

are found for generalist species compared to specialist ones. Second, Condon et al. (2014) showed, using *Drosophila melanogaster*, that specialist and generalists genotypes do not necessary evolve in spatially constant and variable environments, respectively. We hope that our results will stimulate further experimental tests.

Conclusions

Our results show that distinct ecological strategies emerge principally for generalist species, whereas specialist species have much larger ranges of trait values and associations. The drivers of strategies are disturbance, in interaction with species dispersal ability and local recruitment opportunities. In natural habitats, these processes are strongly impacted by anthropogenic changes, such as increased disturbance, habitat fragmentation and habitat loss. This could explain the important changes in the degree of specialisation seen in natural communities (Warren et al. 2001, Munday 2004, Clavel et al. 2010). A better understanding of the characteristics linked to specialisation is therefore essential, as specialists are currently declining worldwide. In particular, the association of specialisation with other life-history traits would benefit from further empirical evidence.

Acknowledgments

The authors thank the editor and three anonymous reviewers, whose comments and suggestions greatly improved the manuscript. We would also like to thank Sergio Rasmann for his improving comments and Kristen Irwin for a careful reading of the manuscript. This work was funded by the Swiss National Science Foundation grants #31003A-112511/2 to LB and PZ00P3–139421/1 and PMPDP3_158381 to SV, and by the FBM Interdisciplinary Grant of the University of Lausanne to SV.

503	
504	
505	References
506	Adler, P. B. et al. 2014. Functional traits explain variation in plant life history strategies. –
507	Proc. Natl. Acad. Sci. USA 111: 740-745.
508	Agrawal, A. A. et al. 2010. Tradeoffs and adaptive negative correlations in evolutionary
509	ecology. Pages 243-268 in M. A. Bell, D. J. Futuyma, W. F. Eanes, and J. S. Levinton,
510	eds. Evolution after darwin: The first 150 years Sinauer, Sunderland, MA, USA.
511	Axmanova, I. et al. 2012. The species richness-productivity relationship in the herb layer of
512	European deciduous forests Global Ecol. Biogeogr. 21:657–667.
513	Bonte, D. et al. 2012. Costs of dispersal. – Biol. Rev. Camb. Phil. Soc. 87: 290–312.
514	Borges, R. M. 2009. Phenotypic plasticity and longevity in plants and animals: Cause and
515	effect? – J. Biosci. 34:605-611.
516	Brown, J. S. and Pavlovic, N. B. 1992. Evolution in heterogeneous environments - effects of
517	migration on habitat specialisation. – Evol. Ecol. 6:360-382.
518	Büchi, L. et al. 2009. The influence of environmental spatial structure on the life-history traits
519	and diversity of species in a metacommunity. – Ecol. Model. 220:2857-2864.
520	Büchi, L. and Vuilleumier, S. 2014. Coexistence of specialist and generalist species is shaped
521	by dispersal and environmental factors. – Am. Nat. 183:612-624.
522	Buoro, M. and Carlson, S. M. 2014. Life-history syndromes: Integrating dispersal through
523	space and time. – Ecol. Lett. 17:756–767.
524	Clavel, J. et al. 2010. Worldwide decline of specialist species: Toward a global functional
525	homogenization? – Front. Ecol. Environ. doi:10.1890/080216.
526	Condon, C. et al. 2014. Temporal variation favors the evolution of generalists in experimental
527	populations of Drosophila melanogaster Evolution 68:720-728.

528	Connell, J. H. 1978. Diversity in tropical rain forests and coral reefs Science 199:1302-
529	1310.
530	Dahirel, M. et al. 2015. Movement propensity and ability correlate with ecological
531	specialization in European land snails: comparative analysis of a dispersal syndrome.
532	– J. Anim. Ecol. 84:228–238.
533	Débarre, F. and Gandon, S. 2010. Evolution of specialization in a spatially continuous
534	environment. – J. Evol. Biol. 23:1090–1099.
535	Devictor, V. and Robert, A. 2009. Measuring community responses to large-scale disturbance
536	in conservation biogeography. – Divers. Distrib. 15:122–130.
537	Devictor, V. et al. 2008. Distribution of specialist and generalist species along spatial
538	gradients of habitat disturbance and fragmentation Oikos 117:507-514.
539	Dynesius, M. and Jansson, R. 2000. Evolutionary consequences of changes in species'
540	geographical distributions driven by Milankovitch climate oscillations Proc. Natl.
541	Acad. Sci. USA 97:9115-9120.
542	Fridley, J. D. et al. 2007. Co-occurrence based assessment of habitat generalists and
543	specialists: a new approach for the measurement of niche width. – J. Ecol. 95:707–
544	722.
545	Futuyma, D. J. and Moreno, G. 1988. The evolution of ecological specialisation. – Annu.
546	Rev. Ecol. Syst. 19:207-233.
547	Gandon, S. and Michalakis, Y. 1999. Evolutionarily stable dispersal rate in a metapopulation
548	with extinctions and kin competition. – J. Theor. Biol. 199:275-290.
549	Goovaerts, P. 1998. Impact of the simulation algorithm magnitude of ergodic fluctuations and
550	number of realizations on the spaces of uncertainty of flow predictions No. 11.
551	Stanford University.

Oikos

552	Grime, J. P. 1977. Evidence for existence of three primary strategies in plants and its
553	relevance to ecological and evolutionary theory. – Am. Nat. 111:1169-1194.
554	Grime, J. P. and Pierce, S. 2012. The evolutionary strategies that shape ecosystems. Wiley-
555	Blackwell.
556	Hastings, A. 1983. Can spatial variation alone lead to selection for dispersal. – Theor. Popul.
557	Biol. 24:244-251.
558	Jakobsson, A., and Eriksson, O. 2000. A comparative study of seed number, seed size,
559	seedling size and recruitment in grassland plants Oikos 88:494-502.
560	Kallimanis, A. S. et al. 2006. Patchy disturbance favours longer dispersal distance. – Evol.
561	Ecol. Res. 8:529-541.
562	Khokhlova, I. S. et al. 2014. A trade-off between quantity and quality of offspring in
563	haematophagous ectoparasites: the effect of the level of specialisation. – J. Anim.
564	Ecol. 83:397–405.
565	Kisdi, E. 2002. Dispersal: Risk spreading versus local adaptation. – Am. Nat. 159:579-596.
566	Klimek, S. et al. 2007. Plant species richness and composition in managed grasslands: The
567	relative importance of field management and environmental factors Biol. Cons.
568	134:559-570.
569	Leishman, M. R. 2001. Does the seed size/number trade-off model determine plant
570	community structure? An assessment of the model mechanisms and their generality
571	Oikos 93:294-302.
572	Levin, S. A. and Muller-Landau, H. C. 2000. The evolution of dispersal and seed size in plant
573	communities. – Evol. Ecol. Res. 2:409-435.
574	Levine, J. M. and Rees, M. 2002. Coexistence and relative abundance in annual plant
575	assemblages: The roles of competition and colonization. – Am. Nat. 160:452-467.
576	Levins, R. 1968. Evolution in changing environments. Princeton Univ. Press.

577	Logue, J. B. et al. 2011. Empirical approaches to metacommunities: a review and comparison
578	with theory Trends Ecol. Evol. 26:482-491.
579	MacArthur, R. H. and Wilson, E. O. 1967. The theory of island biogeography Princeton
580	Univ. Press.
581	Marvier, M. et al. 2004. Habitat destruction, fragmentation, and disturbance promote invasion
582	by habitat generalists in a multispecies metapopulation Risk Anal. 24:869-878.
583	Moles, A. T. and Westoby, M. 2006. Seed size and plant strategy across the whole life cycle.
584	Oikos 113:91-105.
585	Morin, X. et al. 2011. Tree species richness promotes productivity in temperate forests
586	through strong complementarity between species. – Ecol. Lett. 14:1211–1219.
587	Munday, P. L. 2004. Habitat loss, resource specialisation, and extinction on coral reefs. –
588	Glob. Change Biol. 10:1642-1647.
589	Nilsson, A. N. et al. 1993. Macroptery in altitudinal specialists versus brachyptery in
590	generalists - a paradox of alpine scandinavian carabid beetles (coleoptera: Carabidae).
591	– J. Biogeogr. 20:227-234.
592	Nurmi, T. and Parvinen, K. 2008. On the evolution of specialisation with a mechanistic
593	underpinning in structured metapopulations. – Theor. Popul. Biol. 73:222-243.
594	Nurmi, T. and Parvinen, K. 2011. Joint evolution of specialization and dispersal in structured
595	metapopulations. – J. Theor. Biol. 275:78–92.
596	Parvinen, K. and Egas, M. 2004. Dispersal and the evolution of specialisation in a two-habitat
597	type metapopulation. – Theor. Popul. Biol. 66:233-248.
598	Pianka, E. R. 1970. On r-selection and k-selection. – Am. Nat. 104:592-597.
599	Poisot, T. et al. 2011. A conceptual framework for the evolution of ecological specialisation.
600	– Ecol. Lett. 14:841–851.

Oikos

601	Proulx, M. and Mazumder, A. 1998. Reversal of grazing impact on plant species richness in
602	nutrient-poor vs. nutrient-rich ecosystems. – Ecology 79:2581–2592.
603	Ravigné, V. et al. 2009. Live where you thrive: Joint evolution of habitat choice and local
604	adaptation facilitates specialisation and promotes diversity. – Am. Nat. 174:E141-
605	E169.
606	R Core Team 2013. R: A language and environment for statistical computing R Foundation
607	for Statistical Computing, Vienna, Austria.
608	Reich, P. B. 2014. The world-wide 'fast-slow' plant economics spectrum: a traits manifesto. –
609	J. Ecol. 102:275-301.
610	Reich, P. B. et al. 2003. The evolution of plant functional variation: traits, spectra, and
611	strategies Int. J. Plant Sci. 164:S143-S164.
612	Reiff. J. et al. 2015. Linking habitat specialization with species' traits in European birds
613	Oikos
614	Ronce, O. and Kirkpatrick, M. 2001. When sources become sinks: migrational meltdown in
615	heterogeneous habitats Evolution 55:1520–1531.
616	Ronce, O. and Olivieri, I. 1997. Evolution of reproductive effort in a metapopulation with
617	local extinctions and ecological succession. – Am. Nat. 150:220-249.
618	Ronce, O. et al. 2000a. Evolutionarily stable dispersal rates do not always increase with local
619	extinction rates. – Am. Nat. 155:485-496
620	Ronce, O. et al. 2000b. Landscape dynamics and evolution of colonizer syndromes:
621	Interactions between reproductive effort and dispersal in a metapopulation. – Evol.
622	Ecol. 14:233-260.
623	Southwood, T. R. E. 1988. Tactics, strategies and templets Oikos 52:3-18.
624	Stevens, V. M. et al. 2014. A comparative analysis of dispersal syndromes in terrestrial and
625	semi-terrestrial animals. – Ecol. Lett. 17:1039-1052.

626	Tilman, D. 1994. Competition and biodiversity in spatially structured habitats Ecology
627	75:2-16.
628	Venable, D. L. and Brown, J. S. 1988. The selective interactions of dispersal, dormancy, and
629	seed size as adaptations for reducing risk in variable environment Am. Nat.
630	131:360-384.
631	Verberk, W. et al. 2010. Explaining abundance-occupancy relationships in specialists and
632	generalists: A case study on aquatic macroinvertebrates in standing waters. – J. Anim.
633	Ecol. 79:589-601.
634	Vuilleumier, S. et al. 2007. How patch configuration affects the impact of disturbances on
635	metapopulation persistence. – Theor. Popul. Biol. 72:77-85.
636	Warren, M. S. et al. 2001. Rapid responses of british butterflies to opposing forces of climate
637	and habitat change Nature 414:65-69.
638	Westoby, M. 1998. A leaf-height-seed (LHS) plant ecology strategy scheme Plant Soil
639	199:213–227
640	Westoby, M. et al. 1992. Comparative evolutionary ecology of seed size Trends Ecol. Evol.
641	7:368-372.
642	Westoby, M. et al. 2002. Plant ecological strategies: some leading dimensions of variation
643	between species Annu. Rev. Ecol. Syst. 33:125-159.
644	Wright, I. J. et al. 2004. The worldwide leaf economics spectrum Nature 428:821-827.
645	Yu, D. W. and Wilson, H. B. 2001. The competition-colonization trade-off is dead; long live
646	the competition-colonization trade-off. – Am. Nat. 158:49-63.
647	
648	
649	Supplementary material

- 650 Supplementary material (Appendix oik.XXXXX at
- 651 <www.oikosjournal.org/readers/appendix>). Appendix 1-4.



Page 29 of 59 Oikos

Table 1: Values of the parameters used in the simulations. In bold the parameters varying

between simulations.

Parameters	Symbols	Values
Niche optimum	$\mu_{\scriptscriptstyle S}$	random [-2.5 - 2.5]
Niche breadth	$\sigma_{\scriptscriptstyle S}$	random [0.01 - 1]
Competitiveness	k_s	1 ; random* [0.1 - 10]
Reproductive investment	ω_s	1 ; random* [0.1 - 10]
Dispersal ability	δ_s	0.1 ; random [0 - 1]
Survival rate	ψ_s	0.95;0
Disturbance rate	T	0; 0.01; 0.25
*trade-off: $k_s \times \omega_s = 1$		

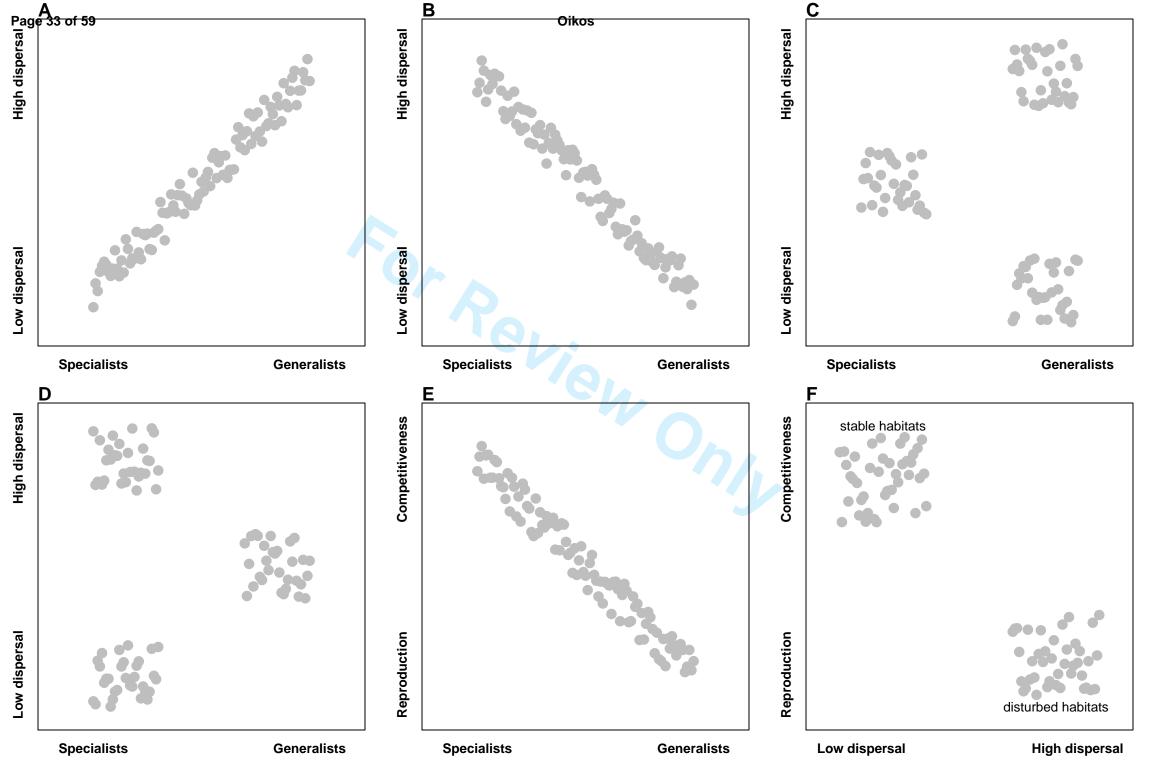
667	Figure legends
668	Figure 1: Expectations for trait strategies from different empirical and theoretical studies. A,
669	B and C: relationship between dispersal and niche breadth, D. relationship between
670	competitiveness (in trade-off with reproductive investment) and niche breadth, E: relationship
671	between competitiveness and dispersal, in stable and disturbed habitats.
672	
673	Figure 2: Association between dispersal and niche breadth values in communities
674	experiencing different strength of disturbance: T = 0: upper row, T = 0.01: middle row, T =
675	0.25: lower row. Results are presented for species with competitiveness and reproductive
676	investment equal to 1, without and with generation overlap (survival rate at each generation is
677	either 0 or 0.95). The grey dots represent the strategies of the surviving species, the black line
678	is the mean dispersal, weighted by species abundances, computed for five categories of niche
679	breadth (0.0-0.1; 0.1-0.2; 0.2-0.4; 0.4-0.6; 0.6-1.0). Each graph represents the pooled results
680	over the 50 replicates.
681	
682	Figure 3: Association between competitiveness (on a logarithmic scale) and niche breadth in
683	communities experiencing different strength of disturbance, $T = 0$: upper row, $T = 0.01$:
684	middle row, and T = 0.25: lower row. Results are presented for species with limited dispersal,
685	without and with generation overlap (survival rate at each generation is either 0 or 0.95). The
686	grey dots represent the strategies of the surviving species, the black line is the mean
687	competitiveness, weighted by species abundances, computed for five categories of niche
688	breadth (0.0-0.1; 0.1-0.2; 0.2-0.4; 0.4-0.6; 0.6-1.0). Each graph represents the pooled results
689	over the 50 replicates.
690	

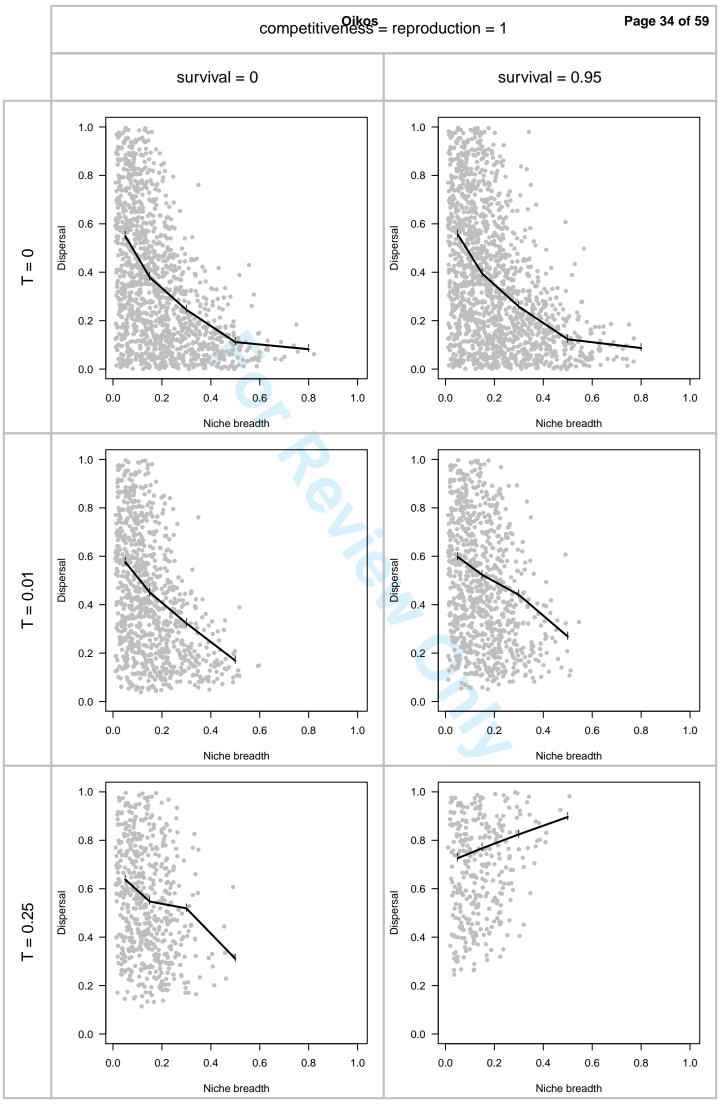
691 Figure 4: Association between dispersal and niche breadth values in communities 692 experiencing different strength of disturbance: T = 0: upper row, T = 0.01: middle row, T = 693 0.25: lower row. Results are presented for species with variable values of competitiveness and 694 reproductive investment, without and with generation overlap (survival rate at each generation 695 is either 0 or 0.95). The grey dots represent the strategies of the surviving species, the black 696 line is the mean dispersal, weighted by species abundances, computed for five categories of 697 niche breadth (0.0-0.1; 0.1-0.2; 0.2-0.4; 0.4-0.6; 0.6-1.0). Each graph represents the pooled 698 results over the 50 replicates. 699 700 Figure 5: Association between competitiveness (on a logarithmic scale) and niche breadth 701 values in communities experiencing different strength of disturbance: T = 0: upper row, in 702 the T = 0.01: middle row, T = 0.25: lower row. Results are presented for species with 703 randomly assigned dispersal values, with and without generation overlap (survival rate at each 704 generation is either 0 or 0.95). The grey dots represent the strategies of the surviving species, 705 the black line is the mean competitiveness, weighted by species abundances, computed for 706 five categories of niche breadth (0.0-0.1; 0.1-0.2; 0.2-0.4; 0.4-0.6; 0.6-1.0). Each graph 707 represents the pooled results over the 50 replicates. 708 709 Figure 6: Association between dispersal ability and competitiveness (log values) for five 710 categories of niche breadth. The mean values are the average over the 50 replicates of the 711 values observed in all surviving individuals. The numbers correspond to the five niche 712 breadth categories from the most specialist species to the most generalist species: 1: 0.0-0.1; 713 2: 0.1-0.2; 3: 0.2-0.4; 4: 0.4-0.6; 5: 0.6-1.0. The cross shows the mean values of the 714 metacommunity at the initial stage of the simulations. Plain lines: no disturbance (T=0),

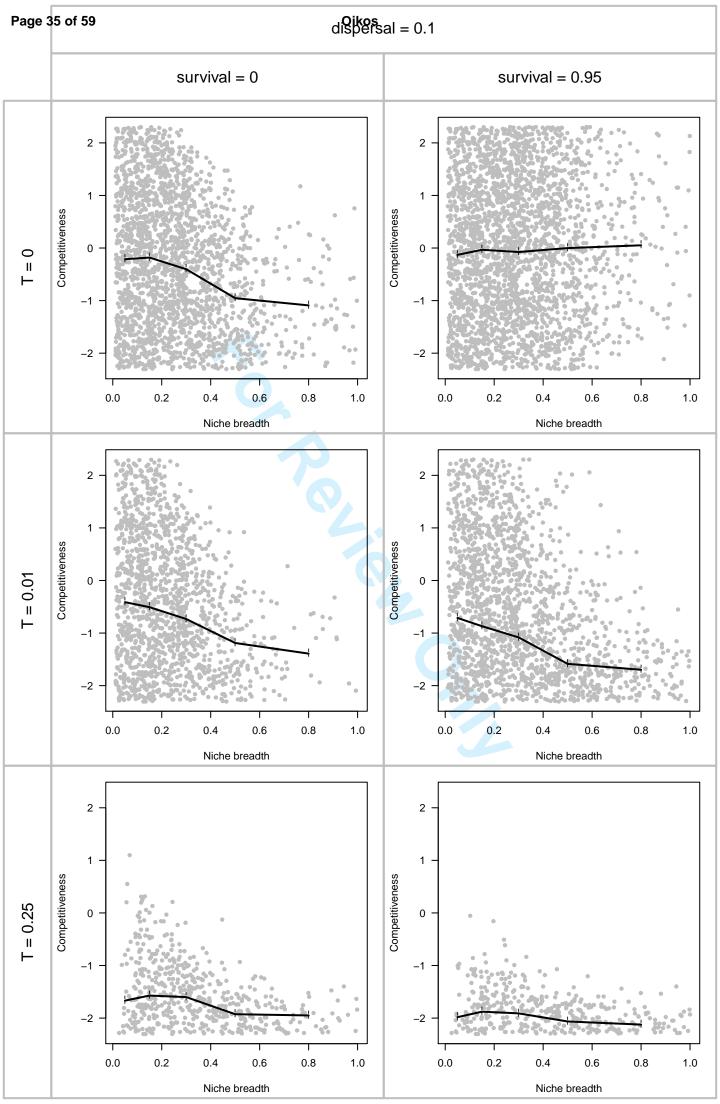
dotted lines: intermediate disturbance rate (T = 0.01), dashed lines: high disturbance rate (T = 0.01)

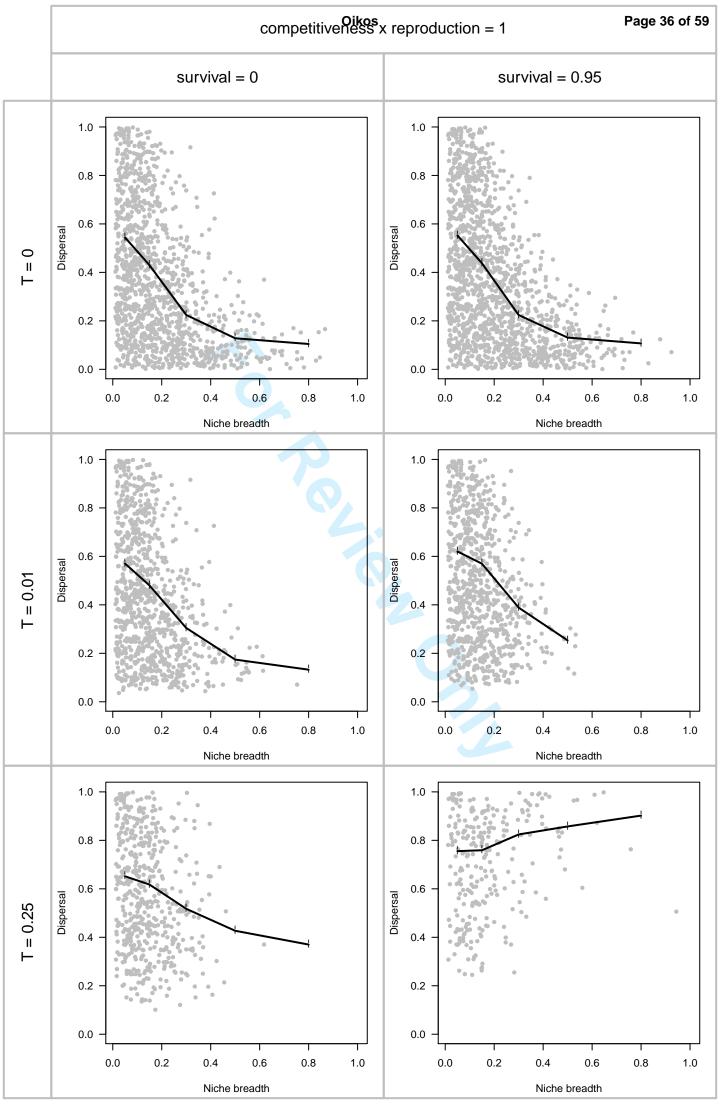
Oikos

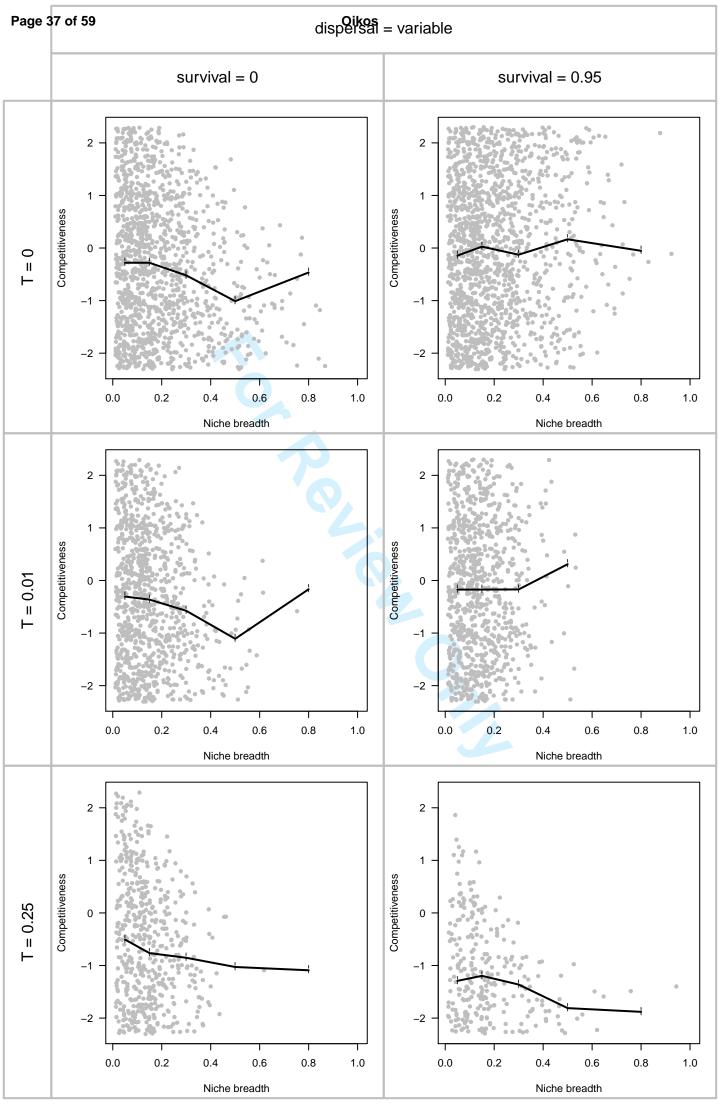
- 716 0.25). In black, non-overlapping generations (survival rate $\psi_s = 0$) and in grey overlapping
- 717 generations (survival rate $\psi_s = 0.95$).

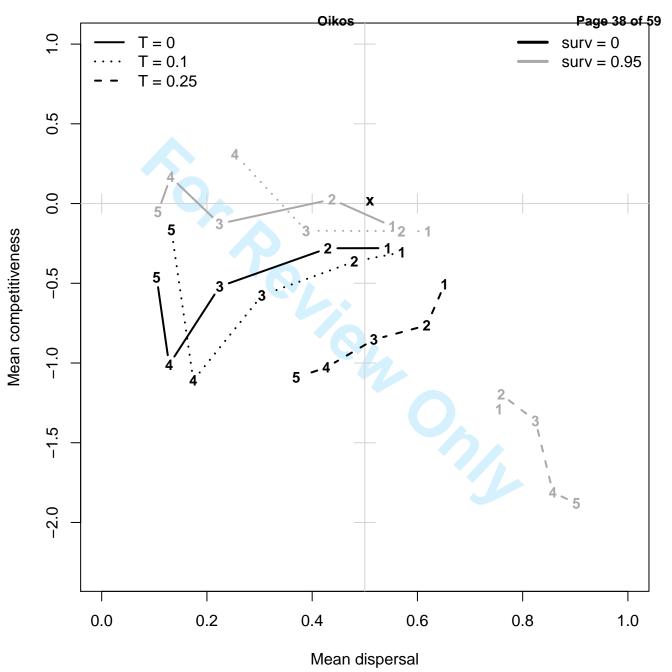












Page 39 of 59 Oikos

Büchi, L. and Vuilleumier, S. 2015 Ecological strategies in stable and disturbed environments depend on species specialisation – Oikos 000:000–000.

Appendix 1: Dynamics of species richness and mean weighted traits

In this Appendix, the trajectories through time of the values of species richness (Figure A1), mean weighted niche breadth (Figure A2), mean weighted competitiveness (Figure A3) and mean weighted dispersal (Figure A4) are represented. In each figure, four panels are displayed. They correspond to the four cases with different values of dispersal ability and generation overlap. A. random dispersal ability δ_s and annual survival rate $\psi_s = 0$, B. random dispersal ability δ_s and annual survival rate $\psi_s = 0.95$, C. dispersal ability $\delta_s = 0.1$ and annual survival rate $\psi_s = 0.95$.

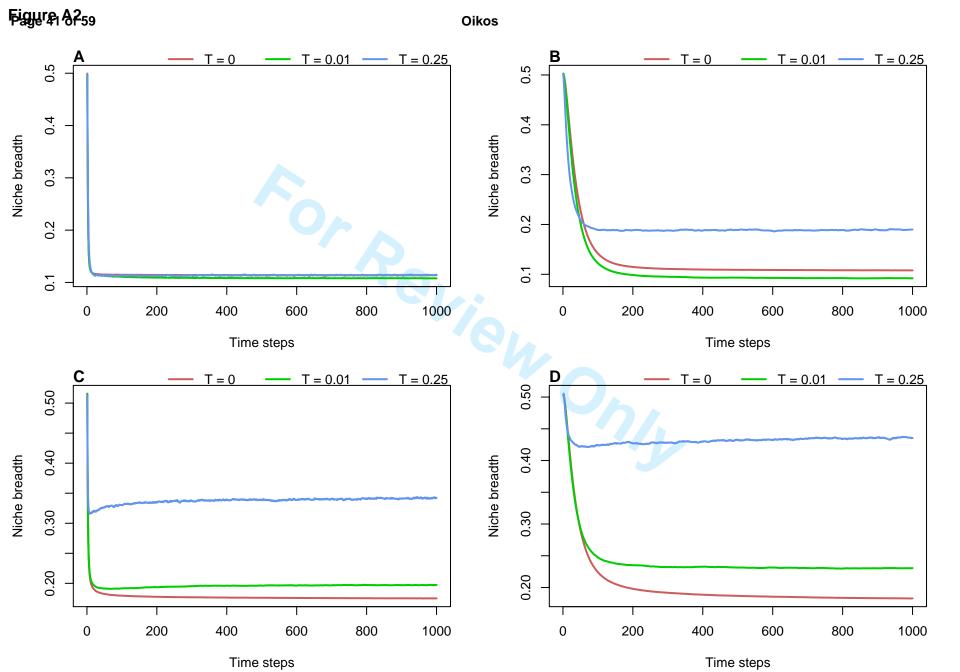
For all panels, the x axis represents the number of time steps (each time step is a full life cycle) and the y axis represents the value of either species richness (Figure A1), mean weighted niche breadth (Figure A2), mean weighted competitiveness (Figure A3) and mean weighted dispersal (Figure A4). In each panel, three lines are represented, the colours of the lines describing different disturbance rate T; red: T = 0, green: T = 0.01, blue: T = 0.25. Each line represents the mean value over 50 replicates.

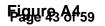
Figure A1: Values of species richness through time.

Figure A2: Values of the mean weighted niche breadth through time.

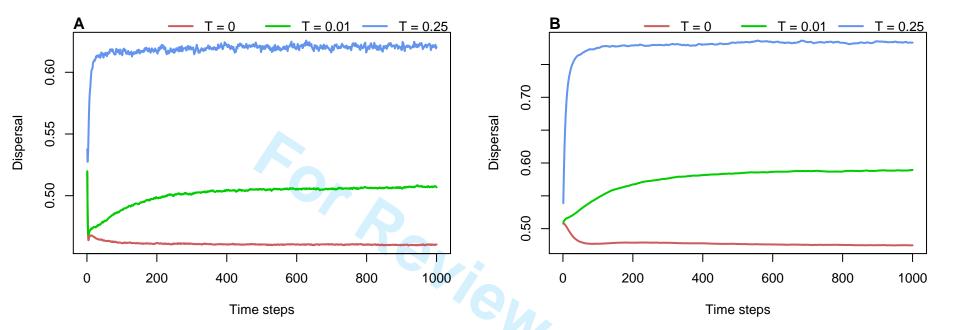
Figure A3: Values of the mean weighted competitiveness through time.

Figure A4: Values of the mean weighted dispersal through time.









Appendix 2: Effect of the number of initial species on the emerging trait strategies

Oikos

Figure A1: Association of dispersal ability with niche breadth, considering values of species competitiveness and reproductive investment each equals to 1 with (A) 10 species at the beginning of the simulations and (B) 1000 species at the beginning of the simulations. The mean values are the average over the 50 replicates of the values observed in all surviving individuals. The dispersal range is the difference between the 97.5% and 2.5% quantiles of the distribution of dispersal values among the surviving species. The numbers correspond to the five niche breadth categories from the most specialist species to the most generalist species: 1: 0.0-0.1; 2: 0.1-0.2; 3: 0.2-0.4; 4: 0.4-0.6; 5: 0.6-1.0. The cross shows the mean and range value of the metacommunity at the initial stage of the simulations. Plain lines: no disturbance (T = 0), dotted lines: intermediate disturbance rate (T = 0.01), dashed lines: high disturbance rate (T = 0.25). In black, non-overlapping generations (survival rate ψ s = 0) and in grey overlapping generations (survival rate ψ s = 0.95).

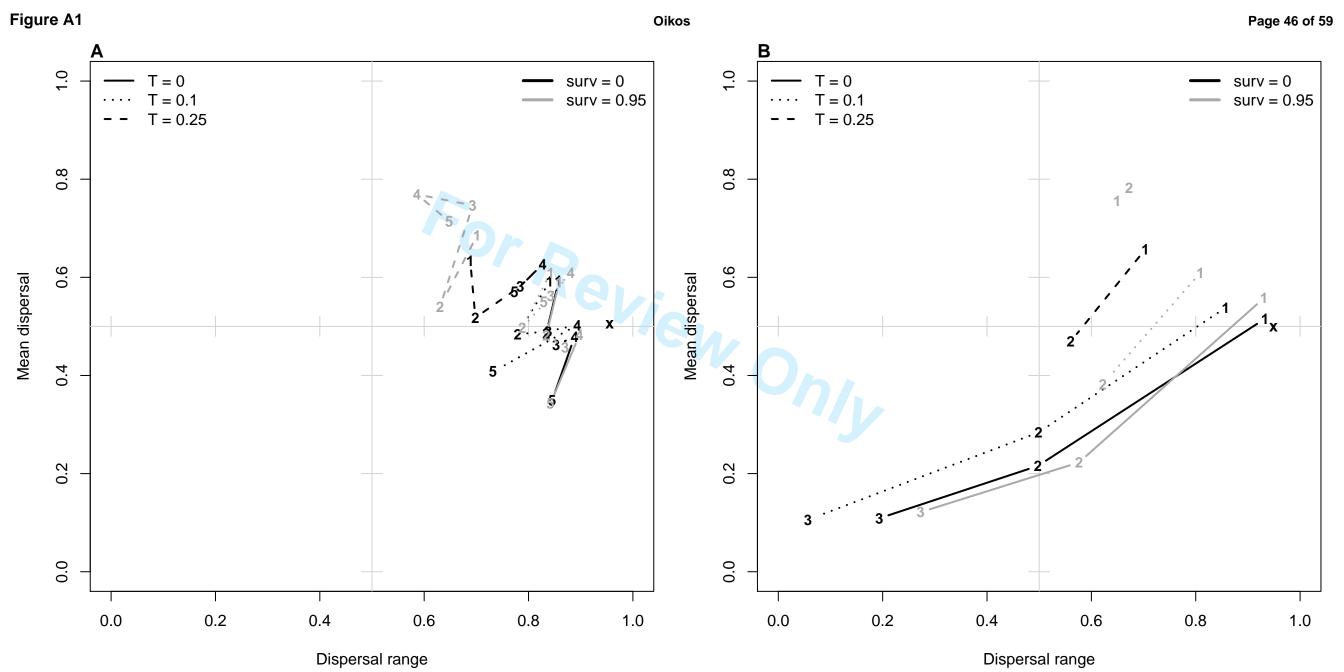
Figure A2: Association between competitiveness (log values) and niche breadth, considering values of dispersal ability equals to 0.1 with (A) 10 species at the beginning of the simulations and (B) 1000 species at the beginning of the simulations. The mean values are the average over the 50 replicates of the values observed in all surviving individuals. The competitiveness range is the difference between the 97.5% and 2.5% quantiles of the distribution of competitiveness values among the surviving species. The numbers correspond to the five niche breadth categories from the most specialist species to the most generalist species: 1: 0.0-0.1; 2: 0.1-0.2; 3: 0.2-0.4; 4: 0.4-0.6; 5: 0.6-1.0. The cross shows the mean and range value of the metacommunity at the initial stage of the simulations. Plain lines: no disturbance (T = 0), dotted lines: intermediate disturbance rate (T = 0.01), dashed lines: high disturbance rate (T = 0.25). In black, non-overlapping generations (survival rate ψ s = 0) and in grey overlapping generations (survival rate ψ s = 0.95).

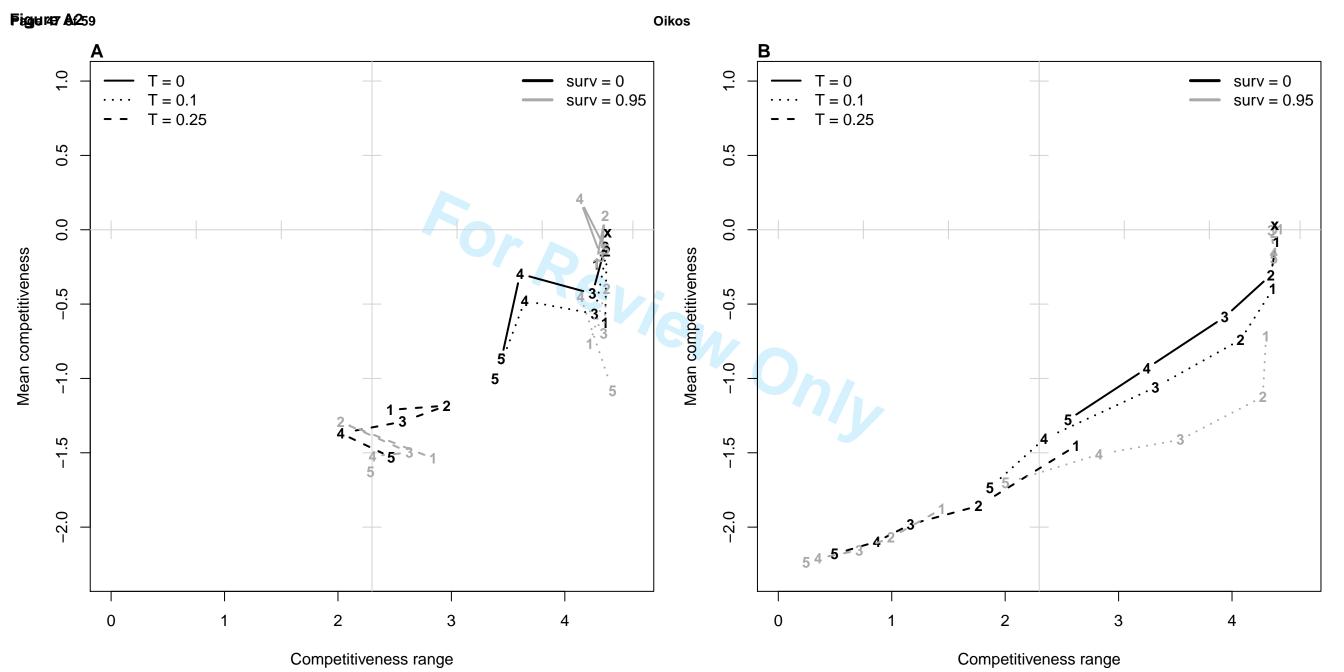
Figure A3: Association between niche breadth and (A, B) dispersal ability (C, D), competitiveness (log values), considering either (A, C) 10 species at the beginning of the simulations or (B, D) 1000 species at the beginning of the simulations. The mean values are the average over the 50 replicates of the values observed in all surviving individuals. The

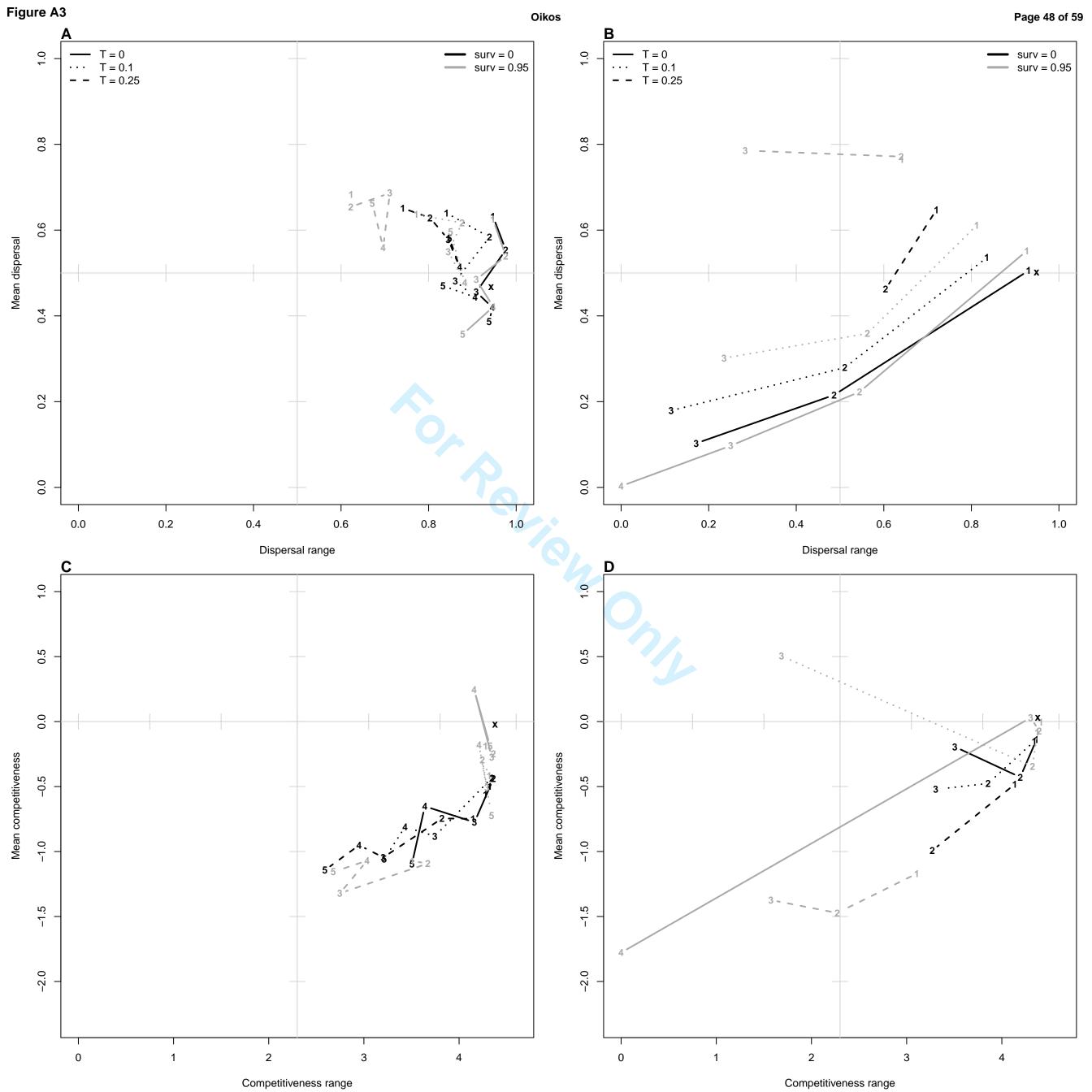
Page 45 of 59 Oikos

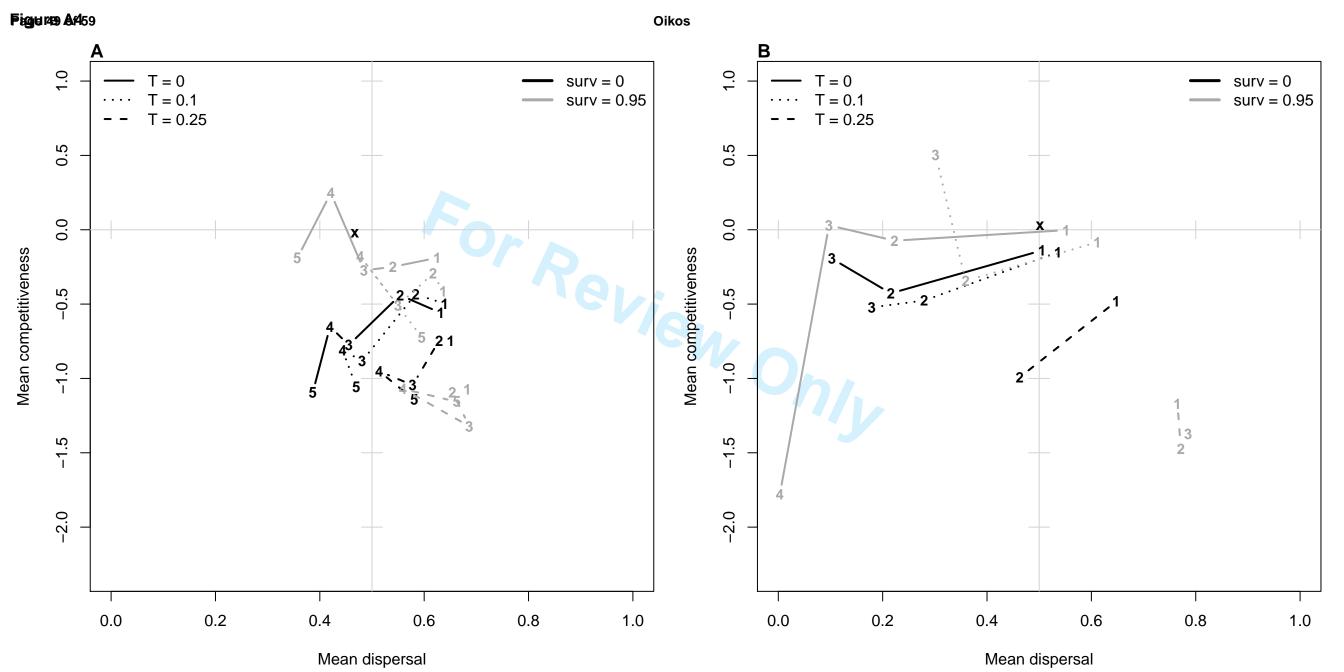
range is the difference between the 97.5% and 2.5% quantiles of distribution of trait values among the surviving species. The numbers correspond to the five niche breadth categories from the most specialist species to the most generalist species: 1: 0.0-0.1; 2: 0.1-0.2; 3: 0.2-0.4; 4: 0.4-0.6; 5: 0.6-1.0. The cross shows the mean and range value of the metacommunity at the initial stage of the simulations. Plain lines: no disturbance (T = 0), dotted lines: intermediate disturbance rate (T = 0.01), dashed lines: high disturbance rate (T = 0.25). In black, non-overlapping generations (survival rate $\psi s = 0.95$).

Figure A4: Association between dispersal ability and competitiveness (log values) for five categories of niche breadth considering either (A) 10 species at the beginning of the simulations or (B) 1000 species at the beginning of the simulations. The mean values are the average over the 50 replicates of the values observed in all surviving individuals. The numbers correspond to the five niche breadth categories from the most specialist species to the most generalist species: 1: 0.0-0.1; 2: 0.1-0.2; 3: 0.2-0.4; 4: 0.4-0.6; 5: 0.6-1.0. The cross shows the mean values of the metacommunity at the initial stage of the simulations. Plain lines: no disturbance (T = 0), dotted lines: intermediate disturbance rate (T = 0.01), dashed lines: high disturbance rate (T = 0.25). In black, non-overlapping generations (survival rate ψs = 0) and in grey overlapping generations (survival rate ψs = 0.95).









Appendix 3: Effect of spatial autocorrelation on emerging trait strategies

Figure A1: Association between dispersal ability and niche breadth, considering species competitiveness and reproductive investment each equals to 1. (A) with moderate spatial autocorrelation ($\alpha = 5$) and (B) with strong spatial autocorrelation ($\alpha = 10$). The mean values are the average over the 50 replicates of the values observed in all surviving individuals. The dispersal range is the difference between the 97.5% and 2.5% quantiles of the distribution of dispersal values among the surviving species. The numbers correspond to the five niche breadth categories from the most specialist species to the most generalist species: 1: 0.0-0.1; 2: 0.1-0.2; 3: 0.2-0.4; 4: 0.4-0.6; 5: 0.6-1.0. The cross shows the mean and range value of the metacommunity at the initial stage of the simulations. Plain lines: no disturbance (T = 0), dotted lines: intermediate disturbance rate (T = 0.01), dashed lines: high disturbance rate (T = 0.25). In black, non-overlapping generations (survival rate ψ s = 0) and in grey overlapping generations (survival rate ψ s = 0.95).

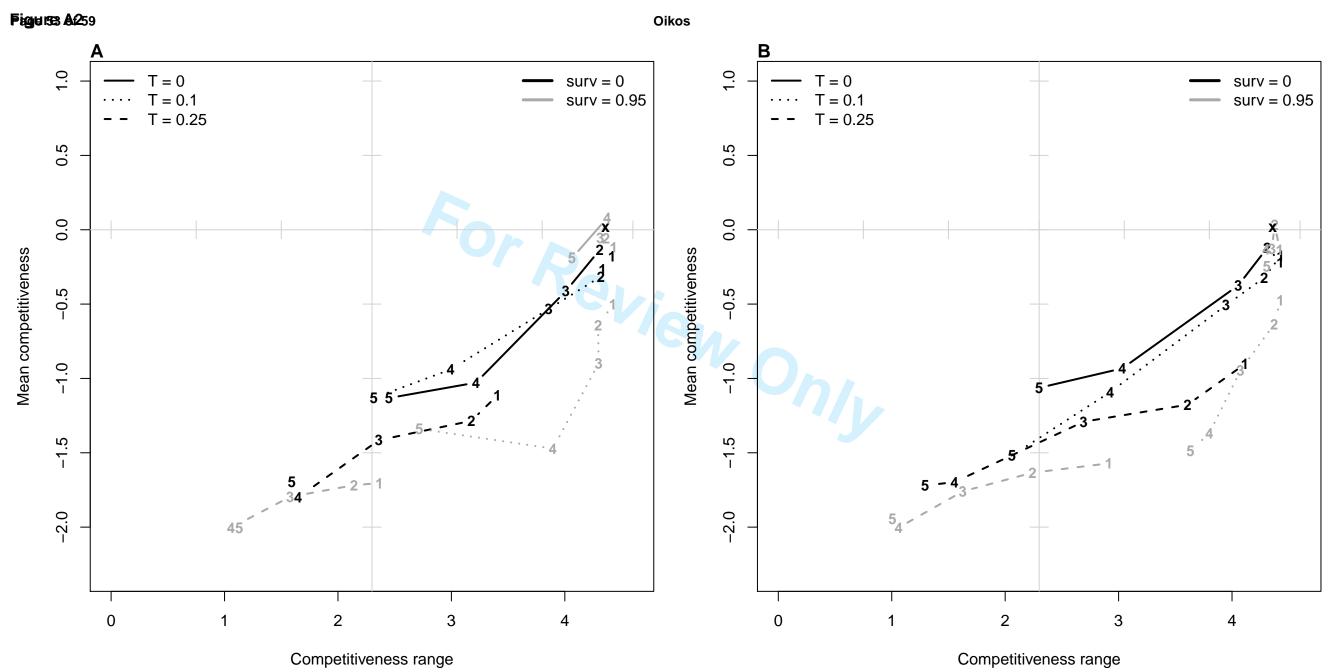
Figure A2: Association between competitiveness (log values) and niche breadth, considering values of dispersal ability equals to 0.1 (A) with moderate spatial autocorrelation ($\alpha = 5$) and (B) with strong spatial autocorrelation ($\alpha = 10$). The mean values are the average over the 50 replicates of the values observed in all surviving individuals. The competitiveness range is the difference between the 97.5% and 2.5% quantiles of the distribution of competitiveness values among the surviving species. The numbers correspond to the five niche breadth categories from the most specialist species to the most generalist species: 1: 0.0-0.1; 2: 0.1-0.2; 3: 0.2-0.4; 4: 0.4-0.6; 5: 0.6-1.0. The cross shows the mean and range value of the metacommunity at the initial stage of the simulations. Plain lines: no disturbance (T = 0), dotted lines: intermediate disturbance rate (T = 0.01), dashed lines: high disturbance rate (T = 0.25). In black, non-overlapping generations (survival rate ψ s = 0) and in grey overlapping generations (survival rate ψ s = 0.95).

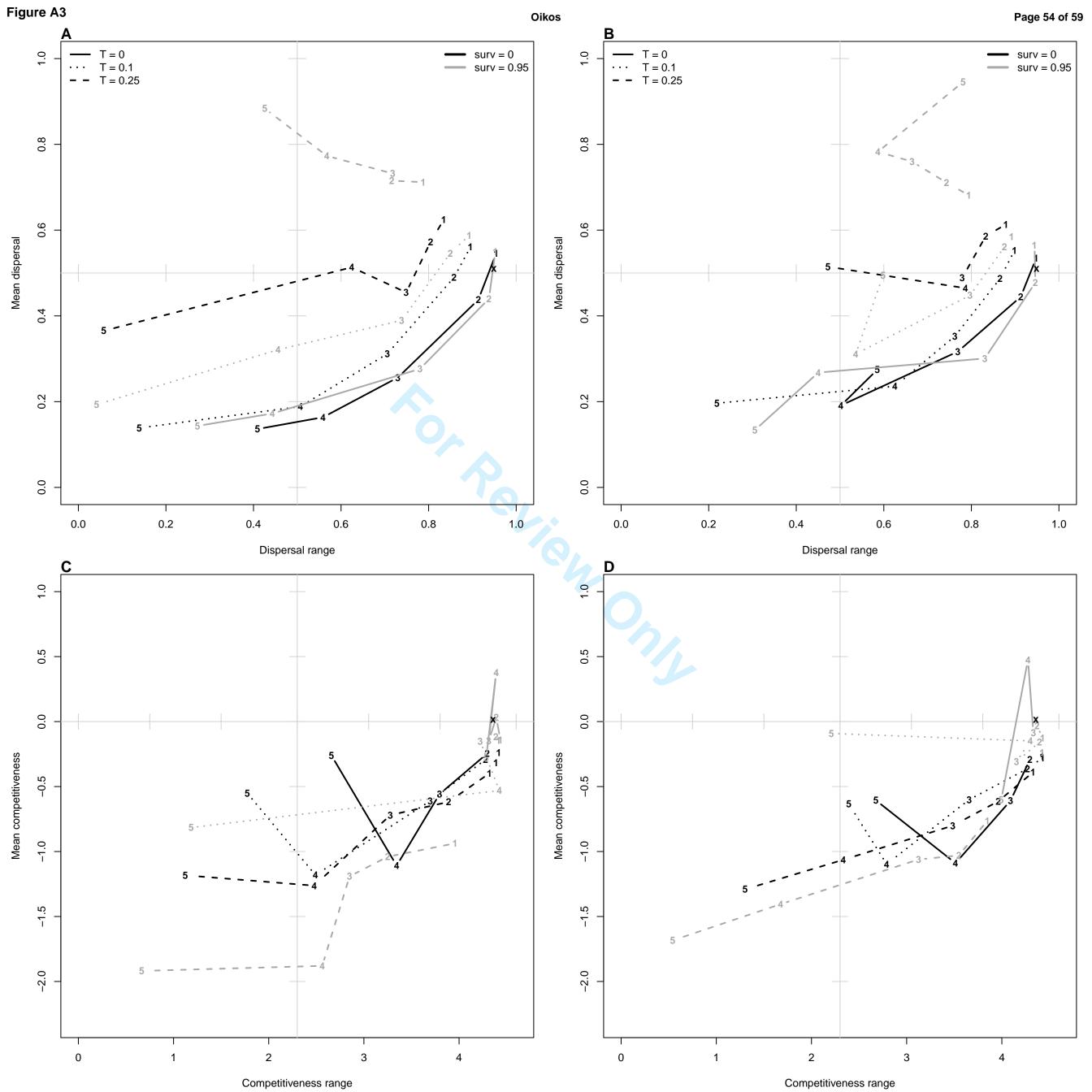
Figure A3: Association between niche breadth and (A, B) dispersal ability, (C, D) competitiveness (log values), considering either (A, C) moderate spatial autocorrelation $(\alpha = 5)$ or (B, D) strong spatial autocorrelation $(\alpha = 10)$. The mean values are the average over the 50 replicates of the values observed in all surviving individuals. The range is the difference between the 97.5% and 2.5% quantiles of distribution of trait values among the surviving

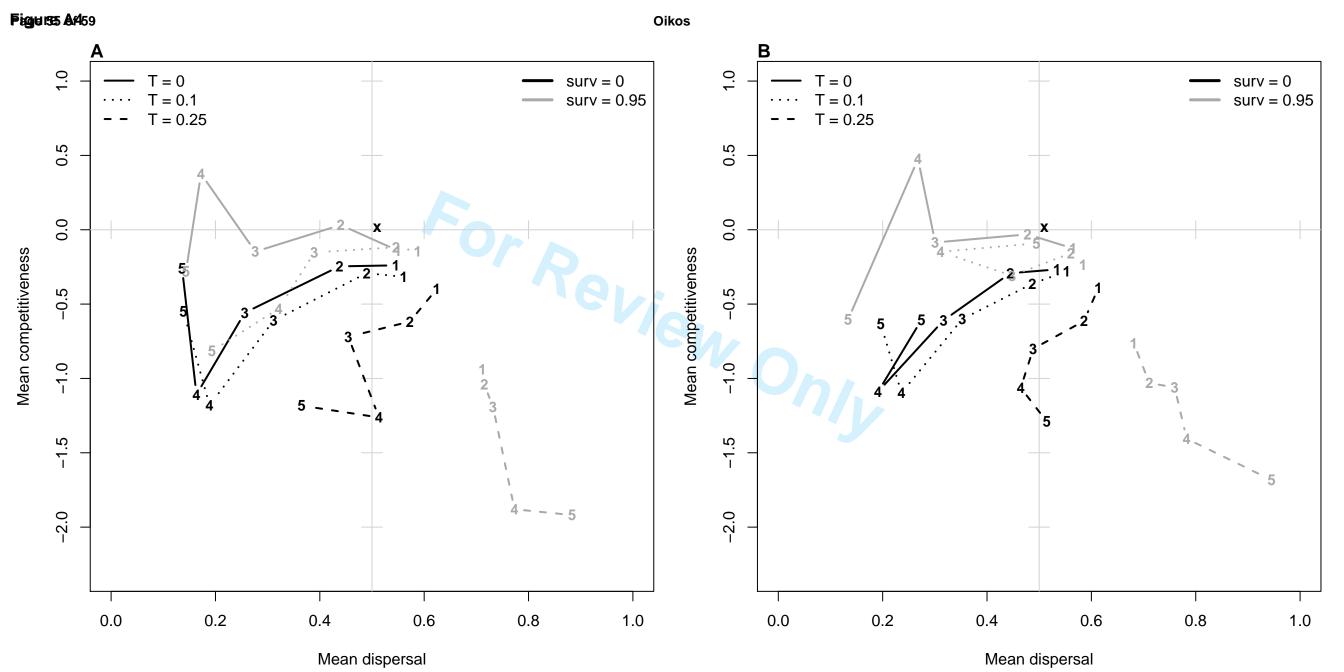
Page 51 of 59 Oikos

species. The numbers correspond to the five niche breadth categories from the most specialist species to the most generalist species: 1: 0.0-0.1; 2: 0.1-0.2; 3: 0.2-0.4; 4: 0.4-0.6; 5: 0.6-1.0. The cross shows the mean and range value of the metacommunity at the initial stage of the simulations. Plain lines: no disturbance (T = 0), dotted lines: intermediate disturbance rate (T = 0.01), dashed lines: high disturbance rate (T = 0.25). In black non-overlapping generations (survival rate ψ s = 0) and in grey overlapping generations (survival rate ψ s = 0.95).

Figure A4: Association between dispersal ability and competitiveness (log values) for five categories of niche breadth considering either (A) moderate spatial autocorrelation ($\alpha = 5$) or (B) strong spatial autocorrelation ($\alpha = 10$). The mean values are the average over the 50 replicates of the values observed in all surviving individuals. The numbers correspond to the five niche breadth categories from the most specialist species to the most generalist species: 1: 0.0-0.1; 2: 0.1-0.2; 3: 0.2-0.4; 4: 0.4-0.6; 5: 0.6-1.0. The cross shows the mean values of the metacommunity at the initial stage of the simulations. Plain lines: no disturbance (T = 0), dotted lines: intermediate disturbance rate (T = 0.01), dashed lines: high disturbance rate (T = 0.25). In black, non-overlapping generations (survival rate ψ s = 0) and in grey overlapping generations (survival rate ψ s = 0.95).







Appendix 4: Relationship between niche breadth and niche optimum

Figure A1: Relationship between niche breadth and niche optimum, with variable dispersal ability, and competitiveness and reproductive investment each equals to 1. (A) without disturbance (T = 0) and non-overlapping generations (survival rate ψ s = 0), (B) with low disturbance rate (T = 0.01) and non-overlapping generations (survival rate ψ s = 0), (C) with high disturbance rate (T = 0.25) and non-overlapping generations (survival rate ψ s = 0), (D) without disturbance (T = 0) and overlapping generations (survival rate ψ s = 0.95), (E) with low disturbance rate (T = 0.01) and overlapping generations (survival rate ψ s = 0.95), (F) with high disturbance rate (T = 0.25) and overlapping generations (survival rate ψ s = 0.95). The size of each point is proportional to the relative abundance of the species.

Figure A2: Relationship between niche breadth and niche optimum, with dispersal ability equals to 0.1, and variable competitiveness and reproductive investment. (A) without disturbance (T = 0) and non-overlapping generations (survival rate ψ s = 0), (B) with low disturbance rate (T = 0.01) and non-overlapping generations (survival rate ψ s = 0), (C) with high disturbance rate (T = 0.25) and non-overlapping generations (survival rate ψ s = 0,9 (D) without disturbance (T = 0) and overlapping generations (survival rate ψ s = 0.95), (E) with low disturbance rate (T = 0.01) and overlapping generations (survival rate ψ s = 0.95), (F) with high disturbance rate (T = 0.25) and overlapping generations (survival rate ψ s = 0.95). The size of each point is proportional to the relative abundance of the species.

Figure A3: Relationship between niche breadth and niche optimum, with variable dispersal ability and species competitiveness and reproductive investment. (A) without disturbance (T = 0) and non-overlapping generations (survival rate $\psi s = 0$), (B) with low disturbance rate (T = 0.01) and non-overlapping generations (survival rate $\psi s = 0$), (C) with high disturbance rate (T = 0.25) and non-overlapping generations (survival rate $\psi s = 0$), (D) without disturbance (T = 0) and overlapping generations (survival rate $\psi s = 0.95$), (E) with low disturbance rate (T = 0.01) and overlapping generations (survival rate $\psi s = 0.95$), (F) with high disturbance rate (T = 0.25) and overlapping generations (survival rate $\psi s = 0.95$). The size of each point is proportional to the relative abundance of the species.

